

1 Running head: Reforestation with native trees in Panama

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6 Responses of 20 native tree species to reforestation strategies for  
7 abandoned farmland in Panama

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1 significantly within four years on badly degraded tropical land (Fisher 1995). Facilitating natural  
2 tree regeneration may be an important management option, but significant barriers to tree  
3 regeneration must be overcome. Lack of seed dispersal is often cited as a major limitation to  
4 natural regeneration (Guevara *et al.* 1986, Janzen 1988, Nepstad *et al.* 1990, Vieira *et al.* 1994,  
5 Aide and Cavelier 1994, Aide *et al.* 1995, da Silva *et al.* 1996), and harsh microsite  
6 characteristics including low nutrient status, high irradiance, high temperature, and low soil  
7 moisture during the dry season can also limit tree seedling survival (Uhl *et al.* 1988, Nepstad  
8 1989, Nepstad *et al.* 1990, Gerhardt 1993).

9         These physiological stresses are worsened when abandoned lands are invaded by exotic  
10 grasses that compete with tree seedlings for moisture and nutrients (Nepstad 1989) and increase  
11 the propensity for fire (D'Antonio and Vitousek 1992). In Panama, abandoned sites in the Canal  
12 area are invaded by the exotic, invasive grass, *Saccharum spontaneum* L. ssp. *spontaneum*.  
13 *Saccharum* attains an average height of 2.5 m and grows in dense, impenetrable stands. It is  
14 drought-adapted, burns frequently, and does not yield to weeding, mulching, fire, or deep  
15 plowing because of its deep rhizomes (Panje 1970). It is listed as one of the most serious weeds  
16 in Indonesia, India, Thailand, the Philippines, and Puerto Rico (Holm *et al.* 1979).

17         Reforestation options in *Saccharum* grasslands have not been studied, but there has been  
18 considerable research on tree restoration of deforested tropical sites invaded by a similar grass  
19 species (*Imperata cylindrica*) in South-East Asia. Nevertheless, no feasible large-scale method  
20 has been found for restoring the original tree cover. Planting indigenous tree seedlings directly  
21 into the *Imperata* had limited success because of the physical difficulty of planting, grass  
22 competition, allelopathy, fire susceptibility, as well as soil degradation and compaction  
23 (Kuusipalo *et al.* 1995). Intensive deep plowing followed by planting of a fast-growing exotic  
24 tree crop to suppress grass and favour natural regeneration has been recommended (Otsamo *et al.*  
25 1995). However, the high cost of site preparation and planting seedlings for this type of  
26 restoration program is economically unattractive for large-scale application in tropical countries.  
27 Instead, forest rehabilitation and management systems are needed that simultaneously accelerate

1 natural regeneration of species-rich native forest while also providing economically and socially  
2 valuable forest products (Parrotta *et al.* 1997). Our goal is to find a low-cost alternative to  
3 intensive mechanical site preparation and tree plantations initiated from nursery-grown seedlings  
4 for restoring productivity of *Saccharum*-dominated grasslands in Panama.

5 If seed source is a major limitation, planting desired species would improve natural  
6 regeneration. Once trees are established, they may act as regeneration nuclei by attracting  
7 vertebrate seed dispersers (Nepstad *et al.* 1990, Lamb *et al.* 1997) and providing favorable  
8 germination and growth microhabitats (Kellman 1985). Native tree species have been  
9 underutilized in tropical reforestation projects, despite their proven ability to grow successfully  
10 on degraded pastures (Gonzalez and Fisher 1994, Butterfield 1995, Butterfield and Espinoza  
11 1995). In this study we planted seeds of twenty native tree species directly into the *Saccharum* to  
12 evaluate their germination, survival, and growth. Species with a range of shade-tolerance and  
13 seed size characteristics were chosen, including small-seeded pioneers and large-seeded, shade-  
14 tolerants, as well as a range in between (Bazazz and Pickett 1980, Augspurger 1984, Denslow  
15 1987, Whitmore 1989). To test the hypothesis that *Saccharum spontaneum* is a major barrier to  
16 native tree establishment, we planted 15,000 seeds of these species in a factorial experiment,  
17 varying above- and below-ground *Saccharum* competition with four shading and mowing  
18 treatments. Two intensities of mowing were used to vary above-ground competition from the  
19 *Saccharum*, leaving roots and thus below-ground competition intact. Two levels of shading (75%  
20 and 95% of full sunlight) were used to decrease below-ground competition from the *Saccharum*  
21 while varying the levels of above-ground competition. Seedling performance may be affected by  
22 proximity to forested areas (Gonzales-Montagut 1996), so we utilized plots at three distances  
23 from the forest.

## 24 MATERIALS AND METHODS

### 25 *Study site*

26 Las Pavas (9°06' N, 79°53' W), Panama, is located 4 km southwest of the Barro Colorado  
27 Nature Monument, where the Smithsonian Tropical Research Institute research reserve is

1 located. The Nature Monument supports tropical moist lowland forest, as did the adjacent study  
2 area prior to deforestation. Deforestation of the public lands between the Canal operating area  
3 and private landholdings began in approximately 1976 (Penna Franco 1990). By 1984 most of  
4 the forest had been cut, burned, and used for subsistence agriculture. Small farm lots were  
5 quickly abandoned, however, most likely due to declining productivity, and were subsequently  
6 invaded by *Saccharum spontaneum*. The entire area has moderately rolling topography, dissected  
7 by small streams that support corridors of native tree species; there are also a few tree islands  
8 100-500 m in diameter in the *Saccharum*.

9         Rainfall on Barro Colorado Island averages 2600 mm, with a pronounced dry season  
10 from mid-December until mid-April (Windsor, 1990). Rainfall was above average during the  
11 first wet season of the study (1996) but well below average during the 1997 dry season  
12 (Smithsonian Environmental Sciences Project, *unpublished data*). Soils are oxisols (Cavelier  
13 1992), with pasture soils characterized by moderate pH (mean 5.97), low total percent nitrogen  
14 (mean 0.34%), and moderate to low available phosphorous (mean 0.0065 mg/g). All other micro  
15 and macronutrients in the upper 10 cm of soil were found in concentrations not limiting to plant  
16 growth; aluminum and iron concentrations were well below toxic levels (Hooper *et al.* in prep.).

### 17                                 *Sampling design*

18         Five sites were chosen to represent a diversity of topographic conditions and were  
19 separated by 0.4 to 0.8 km. Each site was located in an area dominated by *Saccharum*  
20 *spontaneum* adjacent to a different tree island or riparian forest corridor which we refer to as a  
21 forest edge. To determine whether there was an effect of distance from these forest edges on tree  
22 species performance a transect perpendicular to each forest edge was extended into the  
23 *Saccharum*-dominated sites. Plots were placed at 10, 35, and 85 m along this transect. Each plot  
24 consisted of five subplots, each 1 x 8 m in size, for the five treatments: 1) mowing the  
25 *Saccharum* once, 2) mowing three times, 3) mowing three times and shading by 75%, 4) mowing  
26 three times and shading by 95%, 5) control: no mowing or shading. The two shading treatments  
27 were created by construction of a 1.5 m high frame that was placed 0.25 m outside the subplot

1 perimeter and covered with one or two layers of polyurethane shade cloth that removed  
2 approximately 75% (mean 77%, range 73-86%) and 95% (mean 95%, range 93-98%),  
3 respectively, of incident solar radiation. The shade cloth was extended far enough down the sides  
4 of the frame to ensure that each entire subplot was shaded at all times. The subplots were  
5 separated by at least a 1-m pathway which was mown monthly to minimize *Saccharum*  
6 competition from both roots and shading. The shade houses were located 5 m from the three  
7 unshaded treatments so that the latter would not be shaded. *Saccharum* treatments were  
8 implemented by hand-cutting the *Saccharum* with machetes in July 1996, and again in  
9 November 1996 and March 1997 for those mown three times.

#### 10 *Species selection and planting*

11 Twenty species of trees native to Panama were chosen to represent a diversity of families  
12 and shade-tolerance characteristics, and with approximately equal numbers of large-, mid-, and  
13 small-seeded species (Table 1). Species were chosen with the added constraint that, whenever  
14 possible, they have an established ethnobotanical use (Aguilar and Condit 2001). Seeds were  
15 collected from five trees from widely separated locations (A. Sautu, *unpublished data*). From  
16 each tree, 400 seeds were collected, and the seeds from all trees were pooled and mixed.  
17 Exocarps were removed, and seeds were planted within five days of collection for all species not  
18 known to have seed dormancy, and as soon as possible for the rest (Table 1). Seven-hundred-  
19 and-fifty seeds of each species were utilized in this experiment; the rest were utilized to evaluate  
20 each species' light-dependence by measuring germination rates under greenhouse conditions  
21 with varying levels of shade (A. Sautu, *unpublished data*).

22 Ten seeds from each species were planted in each 1 x 8 m subplot, for a total of 200 seeds  
23 per subplot; seeds were located randomly within each of these subplots. Randomization was  
24 introduced by construction of a grid with 20 cm subdivisions; location of each species within this  
25 grid was chosen by random draw. Seeds were planted between July 1996 and March 1997. Four  
26 censuses at two-month intervals (January, March, May, and July 1997) were performed,

1 including height measurements of each seedling. Height measurements were recorded as distance  
2 from the soil surface to the apical meristem with a precision of 1 mm.

3 *Impact of treatments on light, soil moisture, and Saccharum biomass*

4 *Light.*—A Li-Cor photometer (quantum sensor) attached to a Datalogger (Li-Cor,  
5 Lincoln, Nebraska) was used to measure the photon flux density in the photosynthetically active  
6 range (PAR). Instantaneous readings were taken on cloudless days in May 1997 between 10:00  
7 am and 2:00 pm at three heights above the ground: 3 m, 0.5 m (above the *Saccharum* litter), and  
8 0.1 m. Two measurements were taken for each 1 m x 8 m subplot.

9 *Saccharum biomass.*—All live *Saccharum* tissue and all dead litter were collected in a 1 x  
10 1 m quadrat in each subplot at the end of the experiment and also during the third mowing  
11 treatment (March, 1997) to calculate *Saccharum* growth. Samples were dried, and the dry mass  
12 recorded to the nearest gram.

13 *Soil moisture.*— Soil samples were taken from the top 10 cm at the center of each  
14 subplot within a 2-day period in May 1997, 7 days after the previous rainfall. Wet mass was  
15 recorded to the nearest milligram; samples were oven-dried at 60 °C for 2 days, and then re-  
16 weighed. Soil moisture was calculated as 100 (mg H<sub>2</sub>O / mg dry soil).

17 *Statistical Analyses: Overall effects evaluated using ANOVA*

18 A two-way split-plot analysis of variance (ANOVA) was computed using the procedure  
19 GLM in SAS (SAS Institute 1988), with *Saccharum* biomass and light as response variables. The  
20 ANOVA model accounted for distance from the forest with three levels (10 m, 35 m, 85 m) as  
21 the main plot factor, treatment of the *Saccharum* with five levels (once mown, thrice mown,  
22 thrice mown and 75% shaded, thrice mown and 95% shaded, and control) as the sub-plot factor,  
23 and their interaction. Sites were considered a random factor, and we controlled for their effect by  
24 treating sites as statistical blocks. Tukey post-hoc analysis was performed on all variables found  
25 to be significantly ( $P < 0.05$ ) affected by the factors.

26 Repeated-measures ANOVA following the same two-way split-plot design was  
27 computed to determine the effect of distance from the forest, *Saccharum* treatment, time, and

1 their interactions on tree seedling germination, abundance, survival, mean height, and relative  
2 height growth in all unburned sites ( $n = 45$ ). Tree seedling abundance was defined as the number  
3 of individuals alive per time period per subplot. Growth was defined as relative height growth  
4 (RHT), calculated as:

$$5 \quad \text{RHT} = [\ln(H_t) - \ln(H_{t-1})] / T$$

6 where  $H_{t-1}$  and  $H_t$  are measurements of heights on two consecutive dates,  $T$  is the number of  
7 days between these dates, and  $\ln$  is the natural logarithm.

8 *Statistical Analyses: Species-specific effects evaluated using multivariate analyses*

9 *Effect of treatment.* —To understand the effect of treatments on each species, we used  
10 Redundancy Analysis (RDA: Rao 1964) in a MANCOVA-like procedure, as proposed by  
11 Verdonschot and ter Braak (1994). We chose RDA because the species variables in our data  
12 table could not be normalized, making the data unsuitable for standard MANCOVA procedures;  
13 RDA uses permutation testing to find the significance of explained variation and thus species  
14 abundances do not have to be normally distributed prior to analysis. In the present analysis, RDA  
15 was chosen instead of other multivariate procedures which use asymmetrical distance measures  
16 (measures which exclude 0) such as canonical correspondence analysis (CCA), because we  
17 planted a known number of individuals per species, thus making zeros a meaningful basis for  
18 comparison among sites (Legendre and Legendre 1998).

19 RDA was utilized to implement a three-way MANCOVA-like design to test the  
20 significance of the three main factors (distance from the forest, treatment of *Saccharum*, and  
21 time), as well as all two-way interactions, and the three-way interaction. Site and the interaction  
22 of site with all the main factors and interactions were used as covariables in order to treat site as  
23 a statistical block. Each main factor and each interaction was coded as a matrix of predictor  
24 dummy variables by site, much the same as in standard multiple regression with discrete  
25 independent variables except that the dummy variables were made orthogonal so that they would  
26 not be correlated to the covariables (Legendre and Anderson 1999). Analysis involved testing  
27 each main or interaction effect using 999 permutations in the program Canoco 4 (ter Braak and



1 Smilauer 1998), while controlling for the effect of the main factors, sites, and their interactions  
2 not included in a given test. To do this, the matrix of dummy variables by site for the factor of  
3 interest was utilized as the matrix of predictors, and all the other factors and interactions were  
4 placed in another matrix used as covariables in the analysis. This analysis was performed  
5 separately with the following response variable matrices: 1) germination and 2) relative height  
6 growth for all unburned sites ( $n = 45$ ). As with all our multivariate analyses, post-hoc planned  
7 pairwise comparisons were computed using 999 Monte Carlo permutations, and significance  
8 judged after Bonferroni adjustment ( $P < 0.05$  was considered significant).

9 *Effect of fire.*—A wildfire between census two (March) and three (May) burned two out  
10 of the five sites, so all previous analyses were performed on the three unburned plots. To test the  
11 effect of fire on germination and species abundance, a three-way MANOVA-like analysis using  
12 the dummy-variable coding method (detailed above) was performed. The three main factors  
13 included: distance from forest, treatment of the *Saccharum*, and fire (burned, unburned). This  
14 analysis was performed on two species matrices: 1) the number of individuals present for each  
15 species at 30 unburned and 30 burned subplots during the final July census (subplots for one  
16 unburned site were removed from the analysis to achieve a balanced design) and 2) the number  
17 of individuals germinating in these 60 subplots in July 1997.

18 *Survival.*—The effect of the treatments on species survival could not be tested in the same  
19 way as for the germination, abundance, and growth data because for survival data the  
20 symmetrical distance measure used in RDA was not appropriate. This is because, in the data  
21 matrix of survival of species by sites, it is impossible to differentiate a zero resulting from a lack  
22 of germination (inappropriate for evaluating survival) from a zero resulting from death after  
23 germination (appropriate). The survival data did not approach normality, and no transformation  
24 managed to achieve it; thus MANOVA could also not be used. Therefore, we used multiple  
25 regression procedures for each species separately, with survival as the response variable, using  
26 only the data for each site where a given species had germinated. To test the effect of treatment  
27 on survival, a matrix of dummy variables coding for the five treatments was utilized as the

1 predictor variable, and statistical significance was analyzed using permutation testing, which  
2 does not require normality. The partial regression coefficients were used to assess the effect of  
3 the treatments on relative survivorship. To avoid over-specification of the model, the dummy  
4 variables coding for the 95% shaded treatment were removed.

5 *Integrated performance index.*—To assess the overall performance of each species in the  
6 *Saccharum* treatments, an integrated performance index was calculated (De Steven 1991). This  
7 performance index is calculated as the product of germination, survival, and growth. De Steven  
8 (1991) used mean height attained by tree seedlings as a measure of growth whereas we utilized  
9 relative height growth between May and July in order to standardize for differences in timing of  
10 emergence.

11 *Index of light-dependence.*—To evaluate species' light-dependence, relative proportional  
12 germination was calculated per species under greenhouse conditions with 3%, 30% and 50%  
13 incident light levels using seeds collected from the same seed sources (A. Sautu, *unpublished*  
14 *data*). A principal component analysis (PCA) was then conducted on these proportional values,  
15 and the first principal component was utilized as an index of light-dependence (Table 1).

## 16 RESULTS

### 17 *Effect of treatment on Saccharum growth, light levels, and soil moisture*

18 Both shading and mowing treatments significantly ( $F_{[4,24]} = 57.7$ ,  $P < 0.0001$ ) reduced  
19 live *Saccharum* biomass (Table 2). *Saccharum* growth rate was significantly ( $F_{[3,24]} = 17.1$ ,  $P <$   
20  $0.0001$ ) reduced with shading, whereas it did not decrease with mowing ( $P = 0.058$ ; once mown:  
21  $8.9 \pm 1.5 \text{ g m}^{-2} \text{ day}^{-1}$ ; thrice mown:  $10.9 \pm 1.7 \text{ g m}^{-2} \text{ day}^{-1}$ ). Light levels in the photosynthetically  
22 active range (PAR) were significantly affected by *Saccharum* treatment (0.5 m:  $F_{[4,24]} = 15.6$ ,  $P$   
23  $< 0.0001$ ; 0.1 m:  $F_{[4,24]} = 3.2$ ,  $P = 0.029$ ). PAR at 0.5 m above-ground did not significantly differ  
24 when once mown or 75% shaded, averaging 14 -15% of incident solar radiation. PAR also did  
25 not significantly differ between the control and 95% shaded *Saccharum*, averaging 3-5%  
26 incident solar radiation. Soil moisture ranged from  $56.3 \pm 4.2\%$  when 75% shaded and thrice  
27 mown to  $42.0 \pm 3.6\%$  when thrice mown.

1                    *Overall effect of distance from the forest on tree seedling performance*

2                    Neither distance from the forest nor the interaction of distance from the forest with time  
3 or treatment had any significant effect upon tree seedling germination, abundance, survival, or  
4 growth.

5                    *Overall effect of Saccharum treatment on tree seedling performance*

6                    Germination, survival and mean height of tree seedlings were significantly ( $F_{[4,24]} = 11.7$ ,  
7  $P = 0.003$ ;  $F_{[4,24]} = 2.8$ ,  $P = 0.040$ ;  $F_{[4,24]} = 3.9$ ,  $P = 0.004$  respectively) affected by *Saccharum*  
8 treatment; all were higher when the *Saccharum* was artificially shaded in comparison to the  
9 control (Table 3). In contrast, mowing the *Saccharum* three times yearly led to the lowest  
10 germination, survival, and mean height. Percentage germination, survival and mean height nearly  
11 doubled when the *Saccharum* was mown three times and shaded in comparison to when it was  
12 mown three times (germination rose from 13% to 23%, survival from 39% to 79%, and mean  
13 height from 5.5 cm to 9.7 cm). Mowing the *Saccharum* once led to intermediate germination,  
14 survival, and mean height, which were not significantly different from these extremes. Relative  
15 height growth (RHT) was also affected by treatment ( $F_{[4,24]} = 4.9$   $P < 0.010$ ), but this effect was  
16 only significant in the early dry season ( $F_{[4,24]} = 3.5$ ,  $P = 0.025$ ). Height growth was an order of  
17 magnitude higher in both shaded treatments than in the unshaded treatments. In both the  
18 *Saccharum* control and the thrice mown treatment, growth was negative in the dry season,  
19 indicating that increments resulting from growth could not match losses.

20                    The interaction of time and *Saccharum* treatment significantly affected abundance ( $F_{[4,24]}$   
21  $= 13.0$ ,  $P = 0.003$ ; Fig. 1). In January, after the wet season, tree seedling abundance did not differ  
22 among the treatments. In contrast, during the dry season, lower germination and survival in the  
23 unshaded treatments led to a significantly lower abundance of tree seedlings by May in the  
24 unshaded *Saccharum* compared to the shaded. Mean abundance decreased during the early dry  
25 season in both the *Saccharum* control and the unshaded, thrice mown *Saccharum*. At the final  
26 census, in the shaded *Saccharum* (75% and 95%) tree seedling abundance was over double that  
27 found in the unshaded, thrice mown *Saccharum*.

1 Time significantly affected germination ( $F_{[4,24]} = 36.3$ ,  $P < 0.001$ ) and seedling height  
2 ( $F_{[4,24]} = 3.2$ ,  $P = 0.005$ ) (Appendix 1). Germination was higher in the wet seasons (January:  $5.2$   
3  $\pm 1.1\%$ ; July:  $6.4 \pm 0.9\%$ ) than in the dry season (March:  $2.0 \pm 0.4\%$ ; May:  $3.7 \pm 0.8\%$ ). Mean  
4 seedling height was highest in the July wet season ( $9.7 \pm 0.5$  cm) and lowest in the late dry  
5 season (May:  $5.6 \pm 0.5$  cm).

#### 6 *Overall effect of fire*

7 Fire significantly ( $P < 0.001$ ) lowered germination and survival. Following fire, mean  
8 germination was 5.3%, while the mean in all unburned plots was 17.3%. Mean survival (via  
9 resprouting) was 6.7% in burned plots, whereas survival in unburned plots averaged 57.2%.

#### 10 *Species- specific effects*

11 Large-seeded, shade-tolerant species had the highest germination and survival in the  
12 *Saccharum* control (Table 4). Germination correlated positively with seed size ( $P < 0.001$ ):  
13 *Trema* and *Annona*, both with seeds weighing less than 0.1 g, had the lowest mean germination  
14 (3-7%), while *Calophyllum*, which had an average seed size of 14 g, had the highest germination  
15 (42%). Survival also correlated positively with seed size ( $P = 0.002$ ). Survival was so low for  
16 two small-seeded species (*Annona* and *Trema*) that they could not be considered in the remaining  
17 survival analyses. Survival in the *Saccharum* control correlated negatively with the index of  
18 light-dependence ( $P = 0.028$ ), and positively with seed size ( $P = 0.058$ ) suggesting that large-  
19 seeded, shade-tolerant species had higher survival in the uncut *Saccharum*.

#### 20 *Germination*

21 *Effect of Saccharum treatment.* —The germination of all species varied depending upon  
22 *Saccharum* treatment and time (RDA time x treatment interaction:  $F_{[12,24]} = 1.8$ ,  $P = 0.005$ ). The  
23 effect of treatment was only significant during the dry season (March:  $P = 0.001$ ; May  $P =$   
24  $0.002$ ). These differences were pronounced enough that a significant ( $P < 0.001$ ) effect of  
25 *Saccharum* treatment on germination over all time periods was found. This is illustrated by the  
26 ordination of the matrix of overall germination per subplot (Fig. 2). The first axis (horizontal)  
27 explained most of the species variation. It correlated positively with germination in both the 75%

1 and 95% shaded treatments, and negatively with germination in the unshaded treatments. This is  
2 indicated by the position of the treatment vectors in the ordination; both levels of shading are  
3 found on the right side of the ordination, and all the unshaded treatments are located on the left.  
4 Germination did not significantly differ with the level of shading or with the level of mowing.  
5 Thus, species' germination varied depending upon the presence or absence of shading, but was  
6 not affected by mowing.

7         The species most affected by the difference between the shaded and unshaded *Saccharum*  
8 formed two polarized clusters in this ordination space, represented by their positions along the  
9 first ordination axis (Fig. 2). Species in the first cluster had high germination where the  
10 *Saccharum* was shaded, especially *Genipa*, *Ormosia*, *Dipteryx*, *Posoqueria*, and *Heisteria*. The  
11 high correlation between germination of these species and shading is represented in the  
12 ordination by the low angle of their vectors in relation to the first axis, and the high length of  
13 their vectors. They had 15-40% higher germination in shade compared to the control, and  
14 approximately 10% less germination in the thrice mown treatment compared to the control (Fig.  
15 3). The second species group, comprised of *Byrsonima* and *Sterculia* had the opposite response  
16 (Fig. 2); both species had higher germination in the unshaded, thrice mown treatment. The  
17 germination of *Sterculia* doubled, while *Byrsonima* had over 9 times higher germination in the  
18 thrice mown treatment compared to the shaded treatments (Fig. 3). The very small-seeded  
19 species (*Annona*, *Jacaranda*, *Trema* and *Vochysia*) had highest germination in the 75% shaded  
20 treatment (Figs. 2 and 3). Species that germinated in the wet season (*Calophyllum*, *Virola*, and  
21 *Spondias*) were indifferent to the ordination axis, illustrated by their high angle from the first  
22 ordination axis, and the short length of their vectors (Fig. 2).

23         This ordination also suggests that it is not light levels that caused differential germination  
24 in relation to the gradient in shading (Fig. 2). Germination significantly differed between the  
25 uncut *Saccharum* control and the 95% shaded treatment ( $P = 0.014$ ; RDA post-hoc analysis), but  
26 light levels (PAR) did not differ significantly between these treatments. Similarly, germination

1 was significantly different between the 75% shaded and once mown treatment ( $P = 0.011$ ), but  
2 no significant difference existed between PAR in these two treatments.

3 *Effect of fire.*—When burned, most species germinated poorly (RDA;  $P < 0.001$ ).  
4 *Dipteryx*, *Hampea*, *Heisteria*, and *Sterculia* had germination rates approximately cut in half  
5 following fire, while *Posoqueria* and *Genipa* were very sensitive to fire (6.3% germinated on  
6 burned and 42.4% on unburned sites in the former, 7.3% and 39.5% in the latter). *Spondias* had  
7 only slightly lower germination on unburned sites. *Trema* and *Byrsonima* had the opposite  
8 response to fire, with *Trema* averaging 3.6% germination on burned sites and 2.2% on unburned,  
9 and *Byrsonima* 7.6% on burned and 7.0% on unburned sites. Other species could not be  
10 evaluated because they germinated before the wildfire.

11 *Survival*

12 *Effect of Saccharum treatment.*— The survival of all species except *Genipa* was highest  
13 in at least one of the shaded *Saccharum* treatments (Fig. 4). Most species had lower survival in  
14 the thrice mown treatment relative to the control, but *Spondias* and *Sterculia* were opposite.  
15 Survival when the *Saccharum* was mown once was intermediate, and similar to the control for  
16 most species, except *Jacaranda*, *Genipa*, *Spondias* and *Sterculia*, which had over double the  
17 percentage survival when the *Saccharum* was mown once compared to the control.

18 *Effect of fire.*—Species varied in their ability to survive fire. Most died, but five of the  
19 larger-seeded species were able to resprout. Their survival ranged from 20.2% for *Ormosia* to  
20 2.2% for *Calophyllum* and *Virola*; *Carapa* and *Dipteryx* also resprouted.

21 *Growth*

22 *Effect of Saccharum treatment.*— Species varied in their relative growth rate in relation to  
23 treatment and season (RDA time x treatment interaction;  $F_{[4,8]} = 2.9$ ,  $P < 0.001$ ). Ordinations  
24 performed on the species matrix of relative height growth by subplots produced biplots in which  
25 most of the species variation was explained by the first axis (horizontal) in both the wet season  
26 (Fig. 5) and dry season (Fig. 6). In the dry season, the first axis positively correlated with growth  
27 in both the 75% and 95% shaded treatments, but there was no association with mowing. All

1 species in the dry season had higher growth in the shaded treatments (Fig. 5). In the wet season,  
2 both mowing and shading had an effect upon species' growth. The first axis of the ordination  
3 correlated with the presence or absence of shading. Many species with low growth in the control  
4 and thrice mown treatments achieved high growth in the shaded treatments, especially those with  
5 small- to mid-sized seeds (*Jacaranda*, *Posoqueria*, *Heisteria*, *Genipa*, and *Hampea*). A second  
6 axis appears to represent a light gradient: both treatments with the lowest light levels (95%  
7 shaded and control) were positively correlated with this axis, and the treatments with higher light  
8 levels were negatively correlated with this axis. Most species had low growth when the  
9 *Saccharum* was mown three times except *Byrsonima*, *Sterculia*, *Dipteryx* (which achieved their  
10 highest growth rates in this treatment) and *Spondias*.

11 *Overall performance: Integrated performance index*

12 *Ormosia* had the highest integrated performance in the *Saccharum* control, coupling high  
13 growth with moderate survival and germination (Table 5). The large-seeded *Virola*, *Carapa*, and  
14 *Dipteryx* ranked next, with moderate germination and growth, and high survival. *Sterculia* had  
15 relatively high performance in the *Saccharum* control despite low germination because of its  
16 high growth rate. *Calophyllum* followed these with high germination, but low survival and  
17 growth. All other species had relatively poor performance in the *Saccharum* control.

18 Most species had much higher performance where the *Saccharum* was shaded, resulting  
19 from higher germination, survival, and growth, compared to the control. Germination and  
20 survival were similar for most species under both levels of shading, but performance in the 75%  
21 and 95% shaded treatments varied because growth differed. Generally, shade-tolerant species  
22 performed better in the 95% shaded *Saccharum*, whereas *Carapa*, *Dipteryx*, and *Jacaranda* had  
23 higher performance when the *Saccharum* was 75% shaded.

24 For most species, performance was low when the *Saccharum* was mown three times,  
25 except for *Byrsonima*, *Spondias*, and *Sterculia*. *Byrsonima* had zero performance in all other  
26 treatments, indicating its narrow tolerance of shaded conditions. In contrast, *Spondias* performed

1 better when the *Saccharum* was 95% shaded and generally had high performance in all  
2 treatments except the control, indicating its range of tolerance.

### 3 DISCUSSION

4 Our results show that the invasive grass, *Saccharum spontaneum*, negatively affects  
5 germination, survival, and growth of native tree seedlings in abandoned Panamanian farmland.  
6 Most tree species showed a similar response to treatments that reduced *Saccharum* biomass.  
7 Their germination, survival and growth were significantly higher when the *Saccharum* was  
8 shaded in comparison to unshaded control conditions. Shading treatments essentially eliminated  
9 the *Saccharum*, and we presume that below-ground constraints were consequently reduced.  
10 Given that understory light levels were similar when the *Saccharum* was 95% shaded as in the  
11 control, we conclude that below-ground constraints were responsible for the large decrease in  
12 seedling performance when the *Saccharum* was present. In Amazonian pastures, Nepstad (1989)  
13 found that root competition from pasture grasses was a major limitation to tree seedling growth.  
14 He used trenching experiments to ascertain the effect of below-ground competition; we also  
15 recommend trenching experiments to clarify the effect of below-ground competition at our site in  
16 relation to other possible below-ground constraints such as allelopathy or soil factors. But our  
17 objective in the present experiment was to focus on treatments such as mowing and shading  
18 which are easier to implement and thus feasible in local reforestation projects.

19 Mowing the *Saccharum* three times exposed tree seedlings to high irradiance, which  
20 increased temperatures by 5-9°C, decreased humidity 10-15%, and lowered soil moisture by  
21 approximately 10% in comparison to the shaded treatments. We note that these values are  
22 approximate and may differ when larger areas are mown. *Saccharum* growth was undiminished  
23 with mowing. Tree seedling germination, growth, and survival were lowest under these  
24 conditions (approximately half those recorded in the shaded treatments). Other studies of tree  
25 regeneration in neotropical pastures have produced similar results. Gerhardt (1996) found that  
26 germination of *Swietenia macrophylla* was lowest in pasture mown three times yearly in  
27 comparison to unmown pasture, especially during the dry season. Many authors have found that



1 tree seedlings in the neotropics have higher performance under the shade of pasture trees and  
2 shrubs because of protection from irradiation Uhl (1987), higher soil moisture (Kellman and  
3 Miyanishi 1982, Guevara *et al.* 1986, Vieira *et al.* 1994), and decreased root competition with  
4 grasses (Nepstad 1989).

5 We suggest that dry-season water stress, as a result of the dense *Saccharum*, limits tree  
6 seedling regeneration. Elimination of the *Saccharum* through shading significantly improved  
7 most tree species' germination and growth during the dry season. Previous research has shown  
8 that high temperatures and lower moisture availability in tropical pastures, compared to forest,  
9 resulted in water stress and limited seedling establishment and survival. Pasture grasses  
10 significantly decrease soil moisture availability because of their dense root mass in the upper 50  
11 cm of soil (Uhl 1987, Nepstad *et al.* 1990).

12 Although these were the general trends, species were not identical in their response to  
13 mowing and shade, and species differences are important because they offer management  
14 options for different reforestation settings. Because species differences in response to  
15 experimental treatments were consistently predicted by seed size and shade-tolerance, we find it  
16 useful to group species using these characteristics when interpreting our results.

17 We first consider the species with the smallest seeds, *Trema*, *Annona*, *Vochysia* and  
18 *Jacaranda*. They had the lowest overall germination and survival, and their germination was  
19 highest when the *Saccharum* was 75% shaded but much lower with 95% shade, suggesting that  
20 they are light-dependent, as they are on Barro Colorado Island (Welden *et al.* 1991). Survival of  
21 these small-seeded species was very low in the unshaded treatments, suggesting that below-  
22 ground constraints also limit their success in the *Saccharum*. Survival and germination were  
23 lowest during the dry season, suggesting that competition for moisture may also be important.  
24 Nepstad (1989) found that small-seeded, light-demanding species (including *Jacaranda copaia*)  
25 had poor survival in Amazonian pastures because of low soil moisture (Nepstad *et al.* 1990). It  
26 has been suggested that these pioneer species hold the most promise for use in reforestation  
27 because of their fast growth and abundant regeneration in secondary succession (Finegan 1992,

1 Condit *et al.* 1993). In abandoned farmland in Panama, however, they tolerate neither above- nor  
2 below-ground constraints imposed by the *Saccharum* and thus cannot be used.

3 For the next group, we consider three species that had high performance in the exposed  
4 conditions of the thrice mown *Saccharum* treatment and low performance in the *Saccharum*  
5 control. These are the savanna specialist *Byrsonima crassifolia* and the gap-colonists *Spondias*  
6 *mombin* and *Sterculia apetala* (Kellman and Miyanishi 1982, Condit *et al.* 1993). All three were  
7 light-demanding (Index of light-dependence  $> 0.2$ ) and had large seeds (1 to 10 g). Germination  
8 rates for these species were never high, but for *Byrsonima* and *Sterculia* they were highest in the  
9 thrice mown treatment. In contrast to all other species, they had high survival and growth when  
10 the *Saccharum* was mown three times; in fact *Byrsonima* only survived in this treatment, and its  
11 growth rate was over twice that of any other species. We conclude that these species are able to  
12 tolerate dry, exposed conditions as well as below-ground constraints imposed by the *Saccharum*;  
13 their natural regeneration was limited by above-ground constraints, namely light. In addition,  
14 both *Byrsonima* and *Spondias* were able to germinate successfully after fire. Both species are  
15 found frequently in these fire-prone grasslands.

16 Third, we consider a group of three species – *Posoqueria*, *Genipa*, and *Heisteria* – which  
17 had the opposite response to *Saccharum* treatment as the previous light-demanding group. They  
18 had small seeds (0.1 to 1 g) but were shade-tolerant, with high germination in the shaded  
19 treatments and low germination in the mown treatments. They had negative indices of light-  
20 dependence, indicating shade-tolerance as well (Table 1). Since shade did not limit their  
21 performance, we conclude that below-ground constraints resulting from *Saccharum* competition  
22 must have.

23 The large- to very large-seeded (2.9 to 50.4g) species form the final group. They ranged  
24 from the moderately shade-tolerant *Dipteryx* (Index of light-dependence  $-0.4$ , Table 1) to the  
25 very shade-tolerant *Calophyllum*, *Carapa*, and *Virola* (Index of light-dependence  $< -0.9$ ). The  
26 latter three germinated immediately upon planting (in the wet season), irrespective of *Saccharum*  
27 treatment, and despite slow growth rates, they were relatively tall ( $>10$  cm) by the final census.

1 In contrast, *Dipteryx* had dormant seeds and did not germinate until the dry season, grew best in  
2 the treatments with higher light, and showed shade-dependence in germination. All four large-  
3 seeded species had high survival and relatively high performance in the *Saccharum* control  
4 (Table 4). We conclude that the *Saccharum* did not completely limit their regeneration. A  
5 reforestation effort starting with seed and minimal pre-sowing treatment is only likely to succeed  
6 with these large-seeded, shade-tolerant species.

7 Fires burn yearly in the dry season in these *Saccharum*-dominated grasslands, and our  
8 data show that these wildfires are also a major barrier to tree regeneration. Fire killed most  
9 species and significantly lowered the germination of all except *Trema* and *Byrsonima* (but those  
10 two cannot compete with established *Saccharum*). Resprouting from cut stems or stumps is a  
11 very common mechanism for reestablishment following disturbance (Aide *et al.* 1995), and we  
12 found that seedlings of several large-seeded species (*Carapa*, *Dipteryx*, *Virola*, *Ormosia*, and  
13 *Calophyllum*) could resprout following fire. Recurring fires, as a result of grass invasion  
14 following pasture abandonment, arrest natural tree regeneration in abandoned pastures at other  
15 neotropical sites as well (Janzen 1988, Nepstad *et al.* 1991, Aide and Cavelier 1994).

#### 16 *Management suggestions*

17 Fire is a major barrier to tree regeneration at our sites, limiting both establishment and  
18 species diversity. We therefore recommend the establishment of firebreaks, which have also been  
19 an integral part of reforestation strategies in Costa Rica (Janzen 1988) and the Amazon (Nepstad  
20 *et al.* 1990). The breaks must be large for effective fire protection because the flame height of  
21 *Saccharum* wildfires can reach over 15 m.

22 Many alternatives have been suggested for forest restoration throughout the wet tropics.  
23 These range, in order of increasing cost, from simply allowing natural regeneration to proceed, to  
24 planting seeds or seedlings to assist natural regeneration, through establishing tree plantations  
25 and allowing recruitment of tree seedlings below them (Brown and Lugo 1994, Kuusipalo *et al.*  
26 1995, Guariguata *et al.* 1995). Our goal was to find a low-cost strategy for extensive forest  
27 restoration in abandoned Panamanian farmland, and our results suggest that even with the

1 removal of fire, natural tree regeneration will not proceed unassisted because the *Saccharum*  
2 poses a formidable barrier to the small-seeded species that have the highest probability of being  
3 dispersed (eg. Nepstad *et al.* 1996). In general, tree seedling performance was lowest in the  
4 thrice mown treatment, and not significantly different between the unmown and once mown  
5 *Saccharum*. We therefore do not recommend mowing as a site treatment. Shade greatly enhanced  
6 the performance of most tree species and effectively killed the *Saccharum*. Producing a shade  
7 cover as quickly as possible is therefore the best strategy for reforestation. We thus suggest  
8 planting large-seeded tree species (both the moderately and highly shade-tolerant) directly into  
9 the *Saccharum* to catalyze forest regeneration because our data suggest that they can survive in  
10 the *Saccharum*. The shade-tolerant species are advantageous because they have immediate and  
11 high germination, whereas the more light-demanding species have higher growth.

12 In the Brazilian Amazon, Knowles and Parrotta (1995) also found that large-seeded  
13 species could be propagated most efficiently by direct seeding. One limitation of using large-  
14 seeded species, however, is that their seeds generally have limited viability, and must be planted  
15 as soon as possible after collecting. Once a shade cover is produced, we therefore suggest  
16 planting seeds of the smaller-seeded, shade-tolerant species (*Posoqueria*, *Genipa*, and *Heisteria*)  
17 to increase species richness. Their high performance in shaded conditions where the *Saccharum*  
18 was absent suggest that they would perform well, and their longer seed survival and small seed  
19 size are advantageous for collecting and planting. Also, these species could be planted directly  
20 under isolated trees, shrubs, and clumps of *Musa* found in these *Saccharum*-dominated  
21 grasslands, as tried by Vieira *et al.* (1994) in Amazonian pastures.

22 Successful reforestation strategies at other sites such as the Brazilian Amazon (Knowles  
23 and Parrotta 1995) and South-East Asia (Otsamo *et al.* 1995) have involved deep-ripping of the  
24 soil and planting wildlings or nursery-grown seedlings. At our sites, soil trenching would expose  
25 the soil to erosion and introduce high site preparation costs. We do not recommend planting  
26 seedlings directly into the *Saccharum* because it is so dense that it would damage the seedlings.  
27 However, given that firebreaks must be established, we recommend that species of local value be

1 planted as seedlings in the firebreaks. We also recommend direct seeding of the colonizing  
2 species *Byrsonima* and *Spondias* in the firebreaks, where they could benefit from the mowing  
3 treatments undertaken for fire prevention. Both species performed well in dry, exposed  
4 conditions, and both have long-lived seeds which are easy to manipulate and store. Given that  
5 site access is difficult in the *Saccharum*, planting these species in the cleared areas would be  
6 simple and cost-effective. Once established, these species could act as a green firebreak and  
7 attract frugivores, increasing seed dispersal to the regeneration area, as has been found for  
8 windbreaks in Costa Rica (Harvey 2000).

9         Our results suggest that indigenous botanical diversity can help provide a range of cost-  
10 effective reforestation strategies in *Saccharum*-dominated grasslands. A knowledge of species  
11 performance under shaded, exposed, and control conditions is critical for matching the most  
12 promising species to site characteristics, thus maximizing their reforestation potential. Another  
13 barrier that we cannot address is the implementation of reforestation. Demonstration of the  
14 economic and ecological benefits of forest restoration on degraded land is the next necessary  
15 step.

16

17

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## LITERATURE CITED

- 1  
2 Aguilar, S and R. Condit. 2001. Use of native tree species by an hispanic community in Panama.  
3 Economic Botany **55**:223–235.
- 4 Aide, T.M., J.K. Zimmerman, L. Herrera, M. Rosario, and M. Serrano. 1995. Forest recovery in  
5 abandoned tropical pastures in Puerto Rico. Forest Ecology and Management **77**:77–86.
- 6 Aide, T.M., and J. Cavelier. 1994. Barriers to lowland tropical forest restoration in the Sierra  
7 Nevada de Santa Marta, Columbia. Restoration Ecology **2**(4):219–229.
- 8 Augspurger, C.K. 1984. Light requirements of tropical tree seedlings: a comparative study of  
9 growth and survival. Journal of Ecology **72**:777–795.
- 10 Bazazz, F.A., and S.T.A. Pickett. 1980. Physiological ecology of tropical succession: a  
11 comparative review. Annual Review of Ecology and Systematics **11**:287–310.
- 12 Brown, S. and A.E. Lugo. 1994. Rehabilitation of tropical lands: a key to sustaining  
13 development. Restoration Ecology **2** (2-3):97–111.
- 14 Butterfield, R.P. and C.M. Espinoza. 1995. Screening trial of 14 tropical hardwoods with an  
15 emphasis on species native to Costa Rica: fourth year results. New Forests **9**:135–145.
- 16 Butterfield, R.P. 1995. Promoting biodiversity: advances in evaluating native species for  
17 reforestation. Forest Ecology and Management **75**:111–121.
- 18 Cavelier, J. 1992. Fine-root biomass and soil properties in a semideciduous and a lower montane  
19 rain forest in Panama. Plant and Soil **142**:187–201.
- 20 Condit, R., S.P. Hubbell and R.B. Foster. 1993. Identifying fast-growing native trees from the  
21 neotropics using data from a large, permanent census plot. Forest Ecology and  
22 Management **12**:123–143.
- 23 D’Antonio, C.M. and P.M. Vitousek. 1992. Biological invasions by exotic grasses, the grass/fire  
24 cycle, and global change. Annual Review of Ecology and Systematics **23**:63–87.
- 25 da Silva, J.M.C., C. Uhl, and G. Murray. 1996. Plant succession, landscape management, and the  
26 ecology of frugivorous birds in abandoned Amazonian pastures. Conservation Biology  
27 **10**(2):491–503.

- 1 Denslow, J.S. 1987. Tropical rainforest gaps and tree species diversity. Annual Review of  
2 Ecology and Systematics **18**:431–451.
- 3 De Steven, D. 1991. Experiments on mechanisms of tree seedling establishment in old-field  
4 succession: seedling survival and growth. Ecology **72**(3):1076–1088.
- 5 Finegan, B. 1992. The management potential of neotropical secondary lowland rain forest. Forest  
6 Ecology and Management **47**: 295–321.
- 7 Fisher, R.F. 1995. Amelioration of degraded rain forest soils by plantations of native trees. Soil  
8 Science Society of America Journal **59**:544–549.
- 9 Gerhardt, K. 1996. Germination and development of sown mahogany (*Swietenia macrophylla*) in  
10 secondary tropical dry forest habitats in Costa Rica. Journal of Tropical Ecology  
11 **12**:275–284.
- 12 Gerhardt, K. 1993. Tree seedling development in tropical dry abandoned pasture and secondary  
13 forest in Costa Rica. Journal of Vegetation Science **4**:95–102.
- 14 Gonzales, J.E., and R.F. Fisher. 1994. Growth of native forest species planted on abandoned  
15 pasture land in Costa Rica. Forest Ecology and Management **70**:159–167.
- 16 Gonzales-Montagut, R. M. 1996. Establishment of Three Rain Forest Species Along the Riparian  
17 Corridor-Pasture Gradient in Los Tuxtlas, Mexico. Dissertation, Harvard University,  
18 Cambridge, Massachusetts, USA.
- 19 Guariguata, M.R., R. Rheingans, and F. Montagnini. 1995. Early woody invasion under tree  
20 plantations in Costa Rica: Implications for forest restoration. Restoration Ecology **3** (4):  
21 252–260.
- 22 Guevara, S., S.E. Purata, and E. Van der Maarel. 1986. The role of remnant forest trees in  
23 tropical secondary succession. Vegetatio **66**:77–84.
- 24 Harvey, C.A. 2000. Windbreaks enhance seed dispersal into agricultural landscapes in  
25 Monteverde, Costa Rica. Ecological Applications **10** (1):155–173.
- 26 Holm, L., J.V. Pancho, J.P. Herberger, and D.L. Plucknett. 1979. A Geographical Atlas of World  
27 Weeds. John Wiley and Sons Inc., New York, New York, USA.

- 1 Hooper, E., P. Legendre, and R. Condit. Barriers to natural regeneration in abandoned,  
2 deforested land in Panama (in preparation).
- 3 Houghton, R.A., D.S. Lefkowitz, and D.L. Skule. 1991. Changes in the landscape of Latin  
4 America between 1850 and 1985: I Progressive loss of forests. *Forest Ecology and*  
5 *Management* **38**:143–172.
- 6 Janzen, D.H. 1988. Management of habitat fragments in a tropical dry forest: growth. *Annals of*  
7 *the Missouri Botanical Garden* **75**:105–116.
- 8 Kellman, M., and K. Miyanishi. 1982. Forest seedling establishment in neotropical savannas:  
9 observation of experiments in the Mountain Pine Ridge savanna, Belize. *Journal of*  
10 *Biogeography* **9**:193–206.
- 11 Kellman, M. 1985. Soil enrichment by neotropical savanna trees. *Journal of Ecology*  
12 **67**:565–577.
- 13 Knowles, O.H. and J.A. Parrotta. 1995. Amazonian forest restoration: an innovative system for  
14 native species selection based on phenological data and field performance indices.  
15 *Commonwealth Forestry Review* **74** (3): 230-243.
- 16 Kuusipalo, J., A. Goran, J. Yusuf, A. Otsamo, K. Tuomela, and R. Vuokko. 1995. Restoration of  
17 natural vegetation in degraded *Imperata cylindrica* grassland: understory development in  
18 forest plantations. *Journal of Vegetation Science* **6**:205–210.
- 19 Lamb, D., J. Parotta, R. Keenan, and N. Tucker. 1997. Rejoining habitat remnants: restoring  
20 degraded forest lands. *In*: W.F. Laurance and R.O. Bierregaard, Jr. (editors). *Tropical*  
21 *Forest Remnants: Ecology, Management, and Conservation of Fragmented Communities*.  
22 University of Chicago Press, Chicago, Illinois, USA.
- 23 Legendre, P. and M.J. Anderson 1999. Distance-based redundancy analysis: testing multispecies  
24 responses in multifactorial ecological experiments. *Ecological Monographs* **69** (1):1–24.
- 25 Legendre, P. and L. Legendre. 1998. *Numerical Ecology*. 2<sup>nd</sup> English ed. Elsevier Science,  
26 Amsterdam.



- 1 Nepstad, D., C. Uhl, and J.M.C. da Silva. 1996. A comparative study of tree establishment in  
2 abandoned pasture and mature forest in eastern Amazonia. *Oikos* **76**: 25–39.
- 3 Nepstad, D., C. Uhl and E.A. Serrao. 1990. Surmounting barriers to forest regeneration in  
4 abandoned pastures. *In* A. Anderson, editor. *Alternatives to Deforestation: Steps Towards*  
5 *Sustainable Uses of Amazonian Forests*. Columbia University Press, New York, New  
6 York, USA.
- 7 Nepstad, D.C. 1989. *Forest Regrowth in Abandoned Pastures of Eastern Amazonia: Limitations*  
8 *to Tree Seedling Survival and Growth*. Dissertation, Yale University, New Haven,  
9 Connecticut, USA.
- 10 Otsamo, A., G. Adjers, T.S. Hadi, J. Kuusipalo, K. Tuomela, and R. Vuokko. 1995. Effect of site  
11 preparation and initial fertilization on the establishment and growth of four plantation  
12 tree species used in reforestation of *Imperata cylindrica* (L.) Beauv. dominated grasslands.  
13 *Forest Ecology and Management* **73**: 271–277.
- 14 Panje. R.R. 1970. The evolution of a weed. *PANS* **16**(4):590–595.
- 15 Parrotta, J.A., J.W. Turnbull, and N. Jones. 1997. Catalyzing native forest regeneration in  
16 degraded tropical lands. *Forest Ecology and Management* **73**: 271–277.
- 17 Penna Franco, E.A. 1990. *Estudio Socioeconomico y Descripcion de Practicas Agricolas en Tres*  
18 *Comunidades Aledanas al Monumento Natural de Barro Colorado (Las Pavas, Lagartera*  
19 *y Lagarterita)*. Tesis por el titulo de licenciado en ingenieria agronomica. Universidad de  
20 Panama, Panama City, Panama.
- 21 Rao, C.R. 1964. The use and interpretation of principal component analysis in applied research.  
22 *Sankhya A* **26**:329–358.
- 23 SAS Institute Inc. 1988. *Sas/Stat User's Guide*, Release 6.03 Edition. SAS Institute, Carey, NC.
- 24 ter Braak, C.J.F and P. Smilauer. 1998. *Canoco Reference Manual and User's guide to Canoco*  
25 *for Windows: Software for Canonical Community Ordination (version 4)*.  
26 Microcomputer Power. Ithaca New York , USA, 352 pp.

- 1 Uhl, C., R. Buschbacher, and E.A.S. Serrao. 1988. Abandoned pastures in eastern Amazonia. I  
2 Patterns of plant succession. *Journal of Ecology* **76**:663–681.
- 3 Uhl, C. 1987. Factors controlling succession following slash-and-burn agriculture in Amazonia.  
4 *Journal of Ecology* **75**:377–407.
- 5 Vieira, I.G., C. Uhl, and D. Nepstad. 1994. The role of the shrub *Cordia multispicata* Cham. as a  
6 ‘succession facilitator’ in an abandoned pasture, Paragominas, Amazonia. *Vegetatio*  
7 **115**:91–99.
- 8 Verdonschot, P.F.M. and ter Braak, C.J.F. 1994. An experimental manipulation of oligochaete  
9 communities in mesocosms treated with chlorpyrifos or nutrient additions: multivariate  
10 analysis with Monte Carlo permutation tests. *Hydrobiologia* **278**: 251–266.
- 11 Welden, C.W., S.W. Hewett, S.P. Hubbell, and R.B. Foster. 1991. Sapling survival, growth, and  
12 recruitment: relationship to canopy height in a neotropical forest. *Ecology* **72**(1):35–50.
- 13 Whitmore, T.C. 1989. Canopy gaps and the 2 major groups of forest trees. *Ecology* **70**:536–538.
- 14 Windsor, D.M. 1990. Climate and moisture variability in a tropical forest: long-term records  
15 from Barro Colorado Island, Panama. *Smithsonian Contribution to The Earth Sciences*,  
16 no.29. Smithsonian Institution Press, Washington, DC.

TABLE 1. Twenty species of native trees indigenous to Panama planted in the experiment. Throughout the text, species are referred to by their generic names.

Species	Family	Species Code	Seed Mass (g)	Index of Light-Dependence <sup>†</sup>	Dispersal Agent <sup>‡</sup>	Date Collected (mm/dd/yr)	Date Planted (mm/dd/yr)	Germination <sup>††</sup> Date
<i>Annona spraguei</i>	Annonaceae	Annosp	0.04	0.831	animal	09/20/96	09/26/96	1
<i>Antirrhoea trichantha</i>	Rubiaceae	Antitr	0.02		bird	08/02/96	08/18/96	
<i>Byrsonima crassifolia</i>	Malphiaceae	Byrscr	2.08	0.998	animal	07/15/96	08/07/96	3
<i>Calophyllum longifolium</i>	Guttiferae	Calolo	13.70	-0.934	bat	08/27/96	09/01/96	1
<i>Carapa guianensis</i>	Meliaceae	Caragu	50.36	-0.991	animal	10/25/96	10/29/96	1
<i>Ceiba pentandra</i>	Bombacaceae	Ceibpe	0.52		wind	01/10/97	01/15/97	
<i>Dipteryx panamensis</i>	Fabaceae	Diptpa	16.66	-0.379	animal	01/12/97	01/13/97	2
<i>Genipa americana</i>	Rubiaceae	Geniam	0.14	-0.783	animal	02/08/97	02/11/97	2
<i>Hampea appendiculata</i>	Malvaceae	Hampap	0.21	0.146	animal	01/15/97	01/19/97	3
<i>Heisteria concinna</i>	Olacaceae	Heisco	0.30	-0.966	animal	03/06/97	03/08/97	3
<i>Jacaranda copaia</i>	Bignoniaceae	Jac1co	0.01	0.044	wind	09/10/96	09/26/97	1
<i>Lindackeria laurina</i>	Flacourtiaceae	Lindla	0.09		animal	08/01/96	08/07/96	

<i>Ormosia macrocalyx</i>	Fabaceae	Ormoma	0.48	0.619	animal	08/15/96	09/06/96	1
<i>Posoqueria latifolia</i>	Rubiaceae	Posola	0.62	-0.992	animal	01/08/97	01/13/97	2
<i>Spondias mombin</i>	Anacardiaceae	Sponmo	2.18	0.999	animal	08/15/96	08/21/96	1
<i>Sterculia apetala</i>	Sterculiaceae	Sterap	1.28	0.215	animal	01/27/97	01/31/97	3
<i>Trattinickia aspera</i>	Burseraceae	Tratas	0.19		animal	10/28/96	11/23/96	
<i>Trema micrantha</i>	Ulmaceae	Tremmi	0.01	-0.111	bird	08/02/96	09/04/96	1
<i>Virola surinamensis</i>	Myristicaceae	Virosu	2.89	-0.964	animal	07/02/96	07/09/96	1
<i>Vochysia ferruginea</i>	Vochysiaceae	Vocyfe	0.02	0.913	wind	10/10/96	10/14/96	1

*Notes:*

†Index of light-dependence is the first PCA score of proportional germination data under varying light levels in greenhouse conditions (A. Sautu, *unpublished data*); it is left blank for the species that did not germinate during the experiment. Scores ranged from -0.99 (shade-tolerant) to 0.99 (light-demanding).

‡ animal = animal dispersal, wind = wind dispersal, bat = dispersal primarily by bats, bird = dispersal primarily by birds.

††1 = January 1997, 2 = March 1997, 3 = May 1997. Four species did not germinate and are consequently left blank. They did not germinate in the field or in greenhouse trials suggesting that their seed viability was low (A. Sautu, *unpublished data*).

TABLE 1 continued

TABLE 2. Effect of the treatments on *Saccharum*, light (PAR) at 0.5 m above-ground, percent incident solar radiation (I) and our interpretation of above-ground and below-ground constraints imposed by the *Saccharum* under each treatment.

Treatment	Biomass (g)		Growth Rate†		Light‡			Constraints	
	mean	stderr	mean	stderr	mean	stderr	I	above-ground	below-ground
Control	4652 <sup>a</sup>	1134			39 <sup>c</sup>	14	3%	high	high
Once mown	3205 <sup>a</sup>	523	8.9 <sup>a</sup>	1.5	215 <sup>b</sup>	95	15%	low	high
Thrice mown									
Not shaded	1308 <sup>b</sup>	205	10.9 <sup>a</sup>	1.7	550 <sup>a</sup>	89	39%	low	high
75% shaded	179 <sup>c</sup>	60	1.5 <sup>b</sup>	0.5	201 <sup>b</sup>	35	14%	low	low
95% shaded	38 <sup>d</sup>	10	0.3 <sup>c</sup>	0.1	70 <sup>c</sup>	20	5%	high	low

*Notes:*

Similar superscripts indicate no significant ( $P < 0.05$ ) difference according to Tukey post-hoc analysis. *Saccharum* biomass was natural log (ln) transformed, and light levels were  $\ln(x + 0.1)$  transformed.

† grams  $m^{-2}$  day<sup>-1</sup>

‡  $\mu$ moles  $m^{-2}$  sec<sup>-1</sup>

TABLE 3. Effect of *Saccharum* treatment on tree seedling performance: percentage germination, percentage survival, mean seedling height (Height), and dry-season (March-May 1997) relative height growth.

Treatment	% Germination†		% Survival‡		Height (cm)		Relative Height Growth (day <sup>-1</sup> )	
	mean	stderr	mean	stderr	mean	stderr	mean	stderr
Control	14.06 <sup>bc</sup>	1.7	44.23 <sup>b</sup>	9.8	7.34 <sup>b</sup>	0.54	-0.00009 <sup>b</sup>	0.00029
Once mown	17.22 <sup>abc</sup>	1.4	62.30 <sup>ab</sup>	7.6	6.75 <sup>bc</sup>	0.45	0.00027 <sup>b</sup>	0.00027
Thrice mown								
Not shaded	12.95 <sup>c</sup>	1.2	39.00 <sup>b</sup>	6.6	5.47 <sup>c</sup>	0.53	-0.00038 <sup>b</sup>	0.00118
75% shaded	23.11 <sup>a</sup>	1.2	74.13 <sup>a</sup>	5.2	9.70 <sup>a</sup>	0.57	0.00235 <sup>a</sup>	0.00022
95% shaded	19.45 <sup>ab</sup>	1.9	78.92 <sup>a</sup>	5.4	9.14 <sup>a</sup>	0.52	0.00210 <sup>a</sup>	0.00042

*Notes:*

Similar superscripts indicate no significant difference ( $P < 0.05$ ) according to Tukey post-hoc analysis.

† square-root transformed:  $\sqrt{x + 0.01}$

‡ % survival was defined as: (number of individuals that germinated up to and including the May census which survived until the end of the experiment / number of individuals that germinated up to and including the May census)

TABLE 4. Percentage germination (% G) and percentage survival (% S) of all seeds planted in the uncut *Saccharum* control, grouped into 3 seed size classes, and 3 shade-tolerance classes.

Size class	Seed size (g)	Pioneer†		Gap-dependent‡		Shade-tolerant††	
		% G	% S	% G	% S	% G	% S
Small	< 0.15	12.8	35	5.0	0	21.0	0
Medium	0.15 to 1.0			16.1	64	21.1	5
Large	> 1.0	8.3	40	21.1	62	30.7	72

Notes:

† Pioneer (index of light-dependence > 0.75)

‡ Gap-dependent (index of light-dependence 0.75 to -0.75)

†† Shade-tolerant (index of light-dependence < -0.75)

TABLE 5. Integrated performance index (De Steven 1991) for each species, calculated as: (percentage germination x percentage survival x relative height growth). Percentage germination (G), percentage survival (S), Relative height growth x  $10^{-3}$  ( $\text{day}^{-1}$ ) between May and July (RHT) and the performance index (Index) are presented for each species, in each treatment, and listed in descending order in relation to performance in the *Saccharum* control. If no value for survival is given when a germination percentage is listed, that species did not germinate until the final census (July 1997). If a species did not germinate, the columns corresponding to that treatment were left blank. The shaded treatments were also thrice mown.

Species	Control				95% shaded				75% shaded				Thrice mown				Once mown			
	G	S	RHT	Index	G	S	RHT	Index	G	S	RHT	Index	G	S	RHT	Index	G	S	RHT	Index
Ormoma	30	64	3.65	<b>7.0</b>	48	97	4.53	<b>21.1</b>	61	82	3.60	<b>18.0</b>	19	27	3.39	<b>1.7</b>	43	68	2.28	<b>6.6</b>
Caragu	27	77	2.98	<b>6.2</b>	24	95	1.56	<b>3.6</b>	46	79	2.34	<b>8.5</b>	21	74	1.02	<b>1.6</b>	23	86	0.90	<b>1.8</b>
Virosu	24	80	3.05	<b>5.9</b>	26	96	4.79	<b>12.0</b>	20	82	1.26	<b>2.1</b>	19	0		<b>0.0</b>	27	67	0.40	<b>0.7</b>
Diptpa	32	74	2.41	<b>5.7</b>	43	94	2.59	<b>10.5</b>	49	89	2.59	<b>11.3</b>	24	54	6.41	<b>8.3</b>	36	67	4.80	<b>11.6</b>
Sterap	10	50	6.85	<b>3.4</b>	3	100	4.05	<b>1.2</b>	6	100	5.05	<b>3.0</b>	14	60	7.60	<b>6.4</b>	9	100	4.48	<b>4.0</b>
Calolo	41	60	1.07	<b>2.6</b>	41	95	3.79	<b>14.8</b>	36	84	3.73	<b>11.3</b>	43	22	0.16	<b>0.2</b>	49	69	1.59	<b>5.4</b>
Sponmo	14	40	3.33	<b>1.9</b>	27	100	12.85	<b>34.7</b>	24	60	8.02	<b>11.5</b>	24	80	8.34	<b>16.0</b>	34	100	3.20	<b>10.9</b>
Annosp	7	50	4.18	<b>1.5</b>	6	50	3.37	<b>1.0</b>	13	0		<b>0.0</b>	2	0		<b>0.0</b>	8	0		<b>0.0</b>
Vocyfe	19	20	2.90	<b>1.1</b>	17	43	3.52	<b>2.6</b>	20	27	-5.50	<b>-3.0</b>	7	0		<b>0.0</b>	8	25	-2.40	<b>-0.5</b>



Posola	34	50	0.15	<b>0.3</b>	50	63	5.13	<b>16.2</b>	62	64	2.13	<b>8.5</b>	24	40	0.48	<b>0.5</b>	41	25	10.31	<b>10.6</b>
Geniam	21	0		<b>0.0</b>	66	77	6.54	<b>33.2</b>	68	64	6.77	<b>29.5</b>	13	33	1.82	<b>0.8</b>	30	86	8.28	<b>21.4</b>
Jac1co	8	0		<b>0.0</b>	6	40	13.95	<b>3.1</b>	20	50	13.55	<b>13.6</b>	9	0		<b>0.0</b>	9	33	3.66	<b>1.1</b>
Tremmi	2	0		<b>0.0</b>					9	80	5.93	<b>4.3</b>	2	0		<b>0.0</b>	2	0		<b>0.0</b>
Heisco	8				20	100	8.48	<b>17.0</b>												
Byrscr	2												27	63	17.10	<b>28.7</b>				
Hampap	2				9	100	7.42	<b>6.7</b>												

TABLE 5 continued

## Figure captions

**Fig. 1.** Effect of *Saccharum* treatments on mean percentage abundance (abundance / number planted) with the standard error, over all time periods. Similar lower case letters indicate that percentage abundance was not significantly different (Tukey post-hoc analysis,  $P < 0.05$ ).

**Fig. 2.** Ordination biplot illustrating the significant ( $P < 0.001$ ) effect of *Saccharum* treatment on species germination over all time periods. Similar superscripts indicate no significant difference ( $P < 0.05$ ) following post-hoc analysis (planned comparisons). *Saccharum* treatment explained 22.8% of the species variation in germination. Biplot scores are represented following Type II scaling (correlation biplot). Arrows indicate treatments, and lines indicate species vectors. Species codes are listed in Table 1.

**Fig. 3.** Percentage germination in the *Saccharum* treatments vs. control for the 16 species that germinated. Species codes are listed in Table 1.

**Fig. 4.** Percentage seedling survival in the *Saccharum* treatments vs. control for the 11 species we were able to evaluate. Species codes are listed in Table 1. Cases where survival differed significantly ( $P < 0.05$ ) from the 95% shaded treatment are marked with an asterisk. If the asterisk is located on the x axis, without a symbol, survival in the control significantly differed from the 95% shaded treatment.

**Fig. 5.** Ordination biplot illustrating the significant ( $P = 0.002$ ) effect of *Saccharum* treatment on the relative height growth (RHT) of 11 tree species in the dry season, between March and May, 1997 (the other species had not yet germinated). Similar superscripts indicate no significant difference ( $P < 0.05$ ) following post-hoc analysis. *Saccharum* treatment explained 24.1% of the species variation. The first and second axes

explained 15.4% and 5.6% respectively of the species variance. Arrows indicate treatments, and lines indicate species vectors. The codes associated with each species are listed in Table 1 and the symbols representing light requirements and seed sizes match Fig. 2.

**Fig. 6.** Ordination biplot illustrating the significant ( $P < 0.001$ ) effect of *Saccharum* treatment on the relative height growth (RHT) of 16 tree species in the wet season, between May and July, 1997. Similar superscripts indicate no significant difference ( $P < 0.05$ ) based on post-hoc analysis. *Saccharum* treatment explained 19.8% of the variance between species, with the first and second axes explaining 11.2% and 4.5% respectively. Arrows indicate treatments, and lines indicate species vectors. The codes associated with each species are listed in Table 1 and the symbols match Fig. 2.

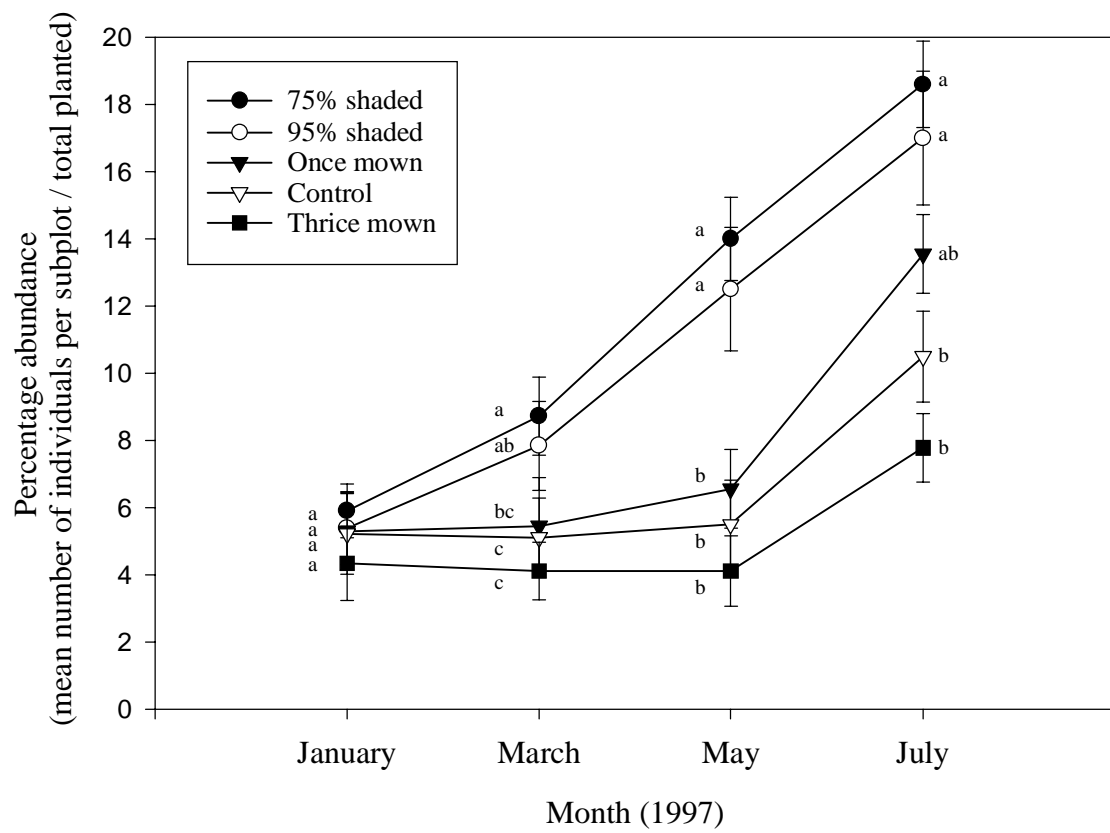
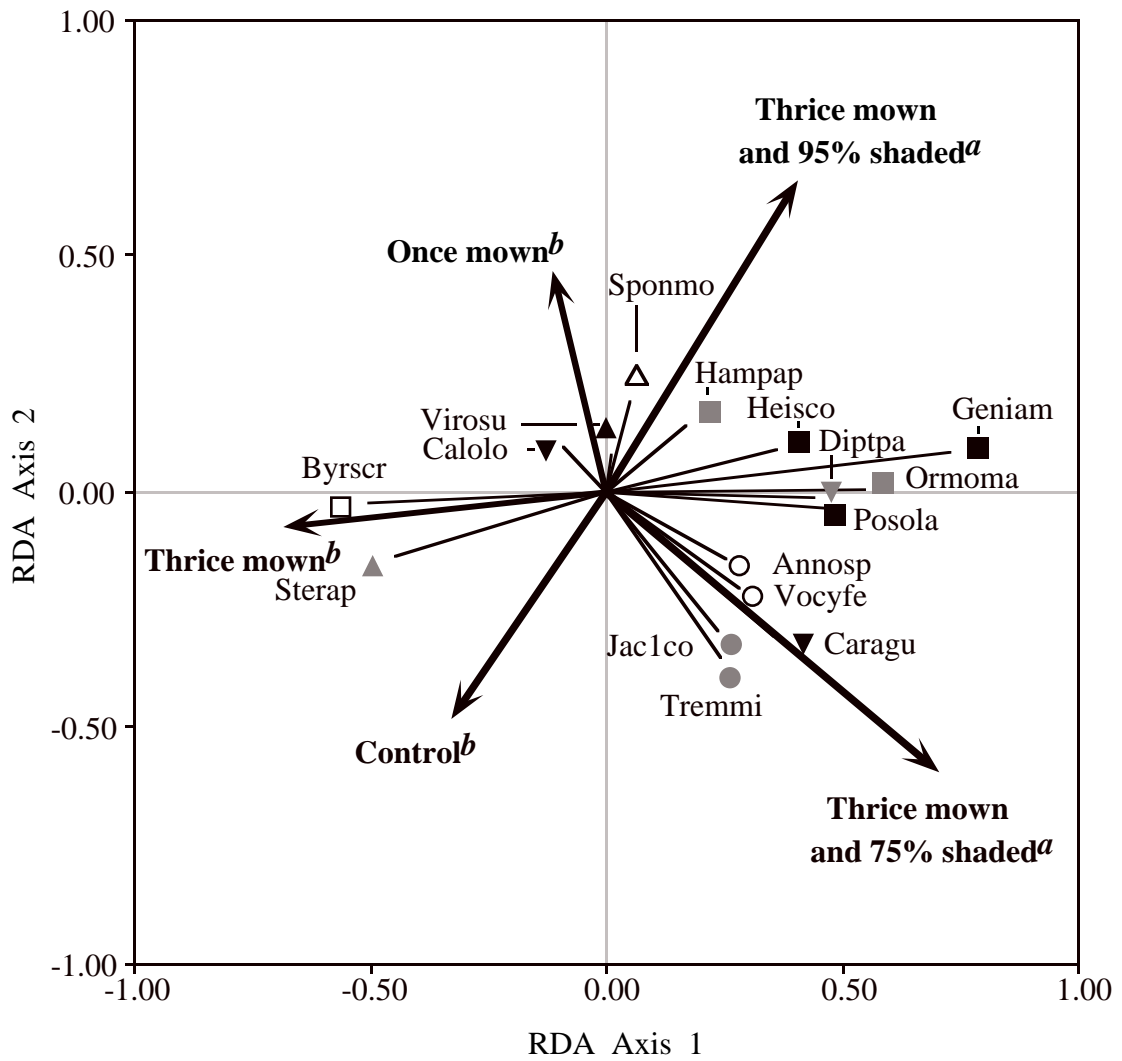


Fig. 1.



- Seed Size:
- = very-small-seeded (0 - 0.1 g)
  - = small-seeded (0.1 - 1 g)
  - ▲ = large-seeded (1 - 10 g)
  - ▼ = very-large-seeded (> 10g)
- Light index:
- white = pioneer / large gap (> 0.75)
  - grey = small gap (0.75 to -0.75)
  - black = light dependence (< -0.75)

Fig. 2.

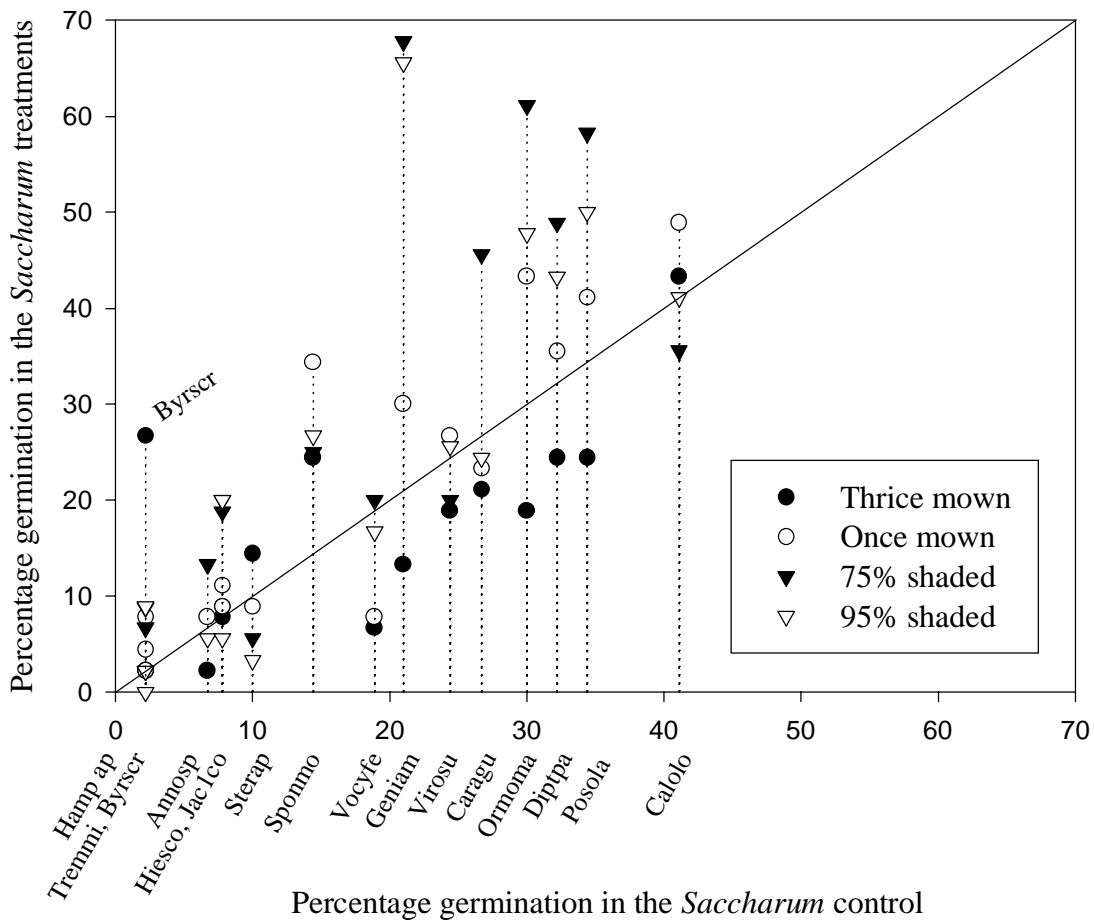
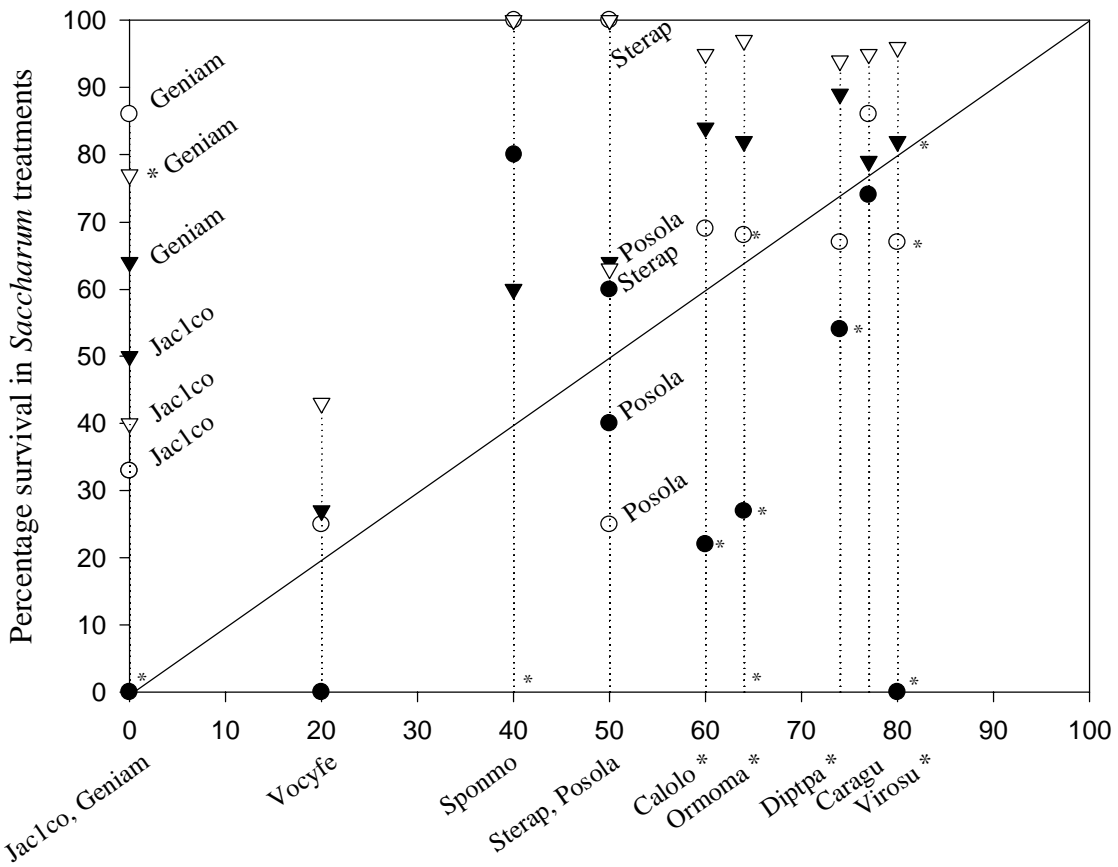


Fig. 3.



Percentage survival in *Saccharum* control

- Thrice mown
- Once mown
- ▼ 75% shaded
- ▽ 95% shaded

Fig. 4.

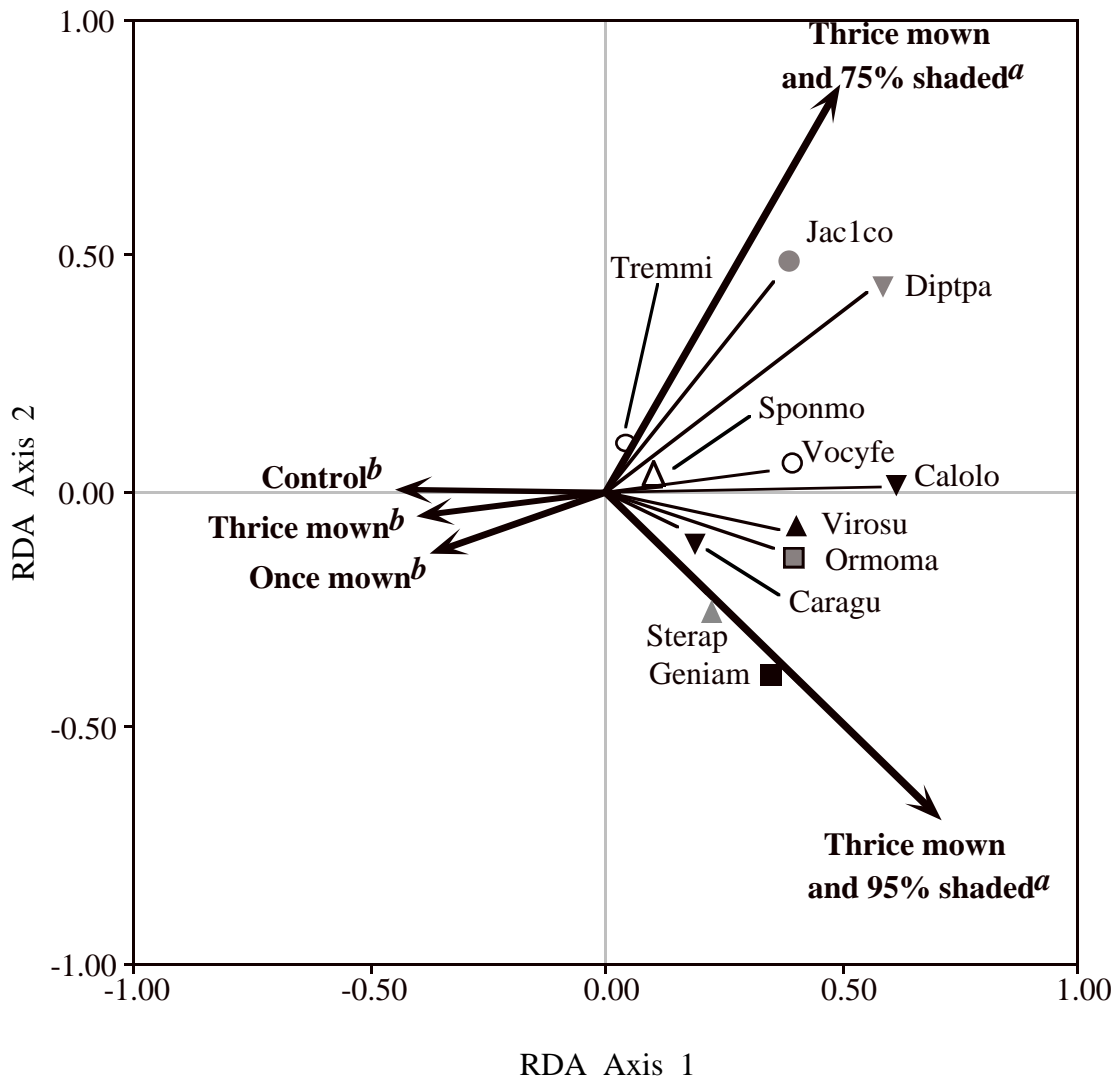


Fig. 5.



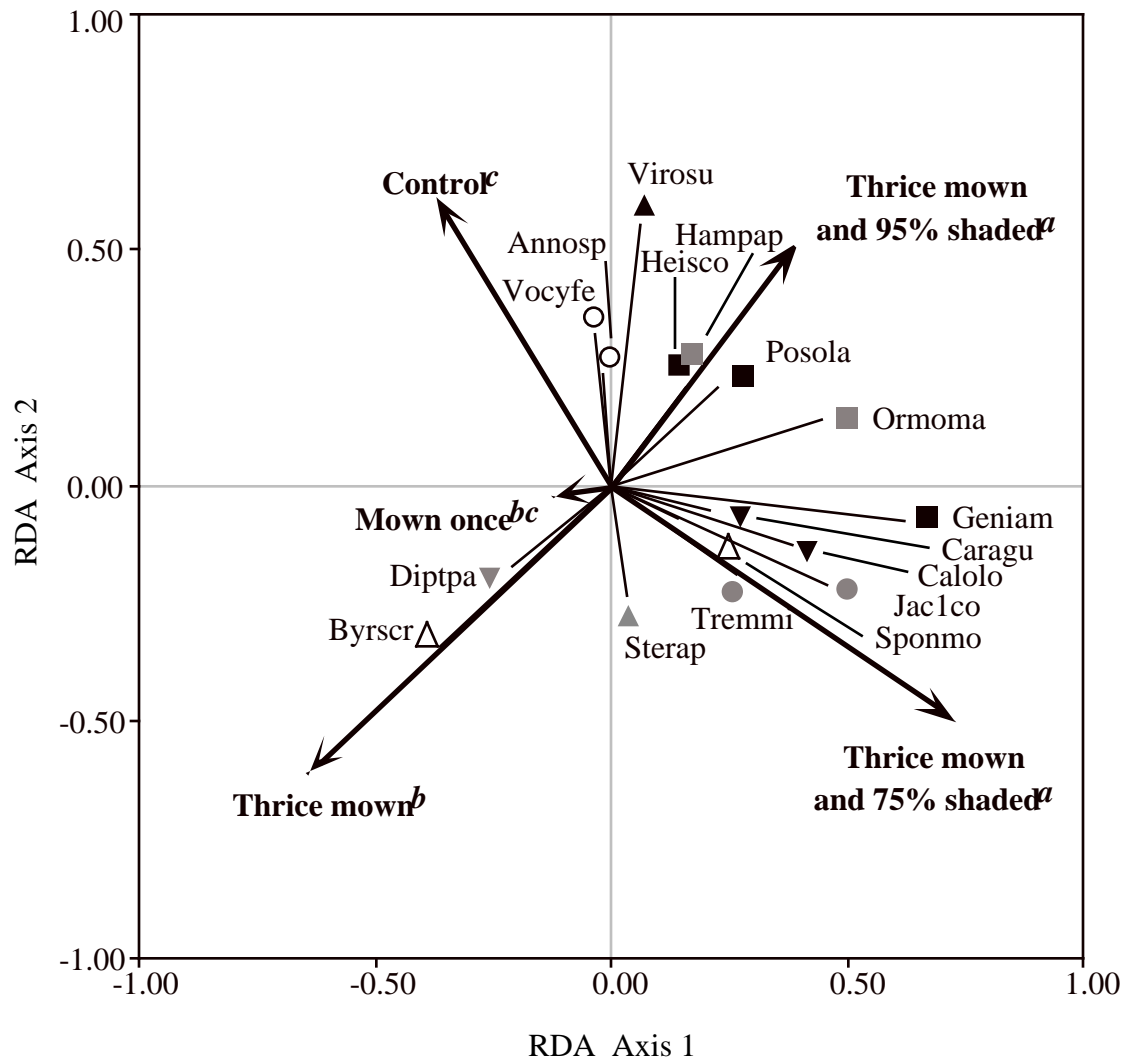


Fig. 6.

APPENDIX 1. Summary of analysis of variance (ANOVA) with repeated measures on % germination, % abundance, % survival, height (cm) and relative height growth  $\times 10^{-3} \text{ day}^{-1}$  (RHT). The analysis followed a repeated measures, split-plot design. Sources of variation included distance from the forest as the main plot factor (distance), shading and mowing treatments of the *Saccharum* as the sub-plot factor (treatment), and their interactions. Site was included as a blocking factor. \*  $P < 0.05$ , \*\*  $P < 0.01$ , \*\*\* $P < 0.005$

Source	Germination†		Abundance†		Survival†		Height‡		RHT‡	
	df	F	df	F	df	F	df	F	df	F
Between subjects										
Distance	2	0.52	2	3.39	2	1.11	2	3.52	2	4.77
Site x distance	4	Main plot error								
Treatment	4	11.68***	4	13.04***	4	2.79*	4	3.22**	4	4.86**
Treatment x distance	8	0.69	8	0.45	8	0.43	8	0.22	8	0.58
Site x treatment + Site x treatment x distance	24	Sub-plot error								

Within subjects										
Time	2	36.31***	3	80.73***	2	1.71	2	5.68**	2	0.86
Time x distance	4	0.59	6	1.01	4	0.15	4	2.17	4	0.13
Time x site x distance	Main plot error									
Time x treatment	8	1.84	12	2.82**	8	1.09	8	1.08	8	0.26
Time x treatment x distance	16	0.92	24	0.80	16	0.74	16	0.36	16	1.08
Time x site x treatment + Time x site x treatment x distance	Sub-plot error									

*Notes:*

† n = 45 subjects (i.e. 45 unburned subplots); ‡ n = 371 subjects (i.e. 371 seedlings were present over all 3 time periods in the 45 unburned subplots)

APPENDIX 1 continued

APPENDIX 2. Summary of RDA in a MANCOVA-like design on matrices of germination per species per sub-plot per time period (Germination) or relative growth rates for height  $\times 10^{-3} \text{ day}^{-1}$  per species per subplot per time period (RHT). Sources of variation included distance from the forest (distance), shading and mowing treatments of the *Saccharum* (treatment), time and their interactions. Site and the interaction of site with all main factors and interactions were utilized as covariables. \*  $P < 0.05$ , \*\*  $P < 0.01$ , \*\*\* $P < 0.005$  (determined using permutation testing).

Source	Germination		RHT	
	df	F	df	F
Site	2		2	
Distance	2	1.03	2	1.41
Treatment	4	2.89**	4	4.67***
Treatment x distance	8	0.70	8	1.25
Time	3	31.51**	1	12.08**
Time x distance	6	1.01	2	2.46*
Time x treatment	12	1.81**	4	2.85***
Time x treatment x distance	24	0.72	8	1.36