Water Treatment and Distribution System Improvements for Mae La Refugee Camp, Thailand

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1 INTRODUCTION

Among the 1.1 billion people worldwide without UN-qualified improved drinking water, refugees are a population with unique needs (UNICEF & WHO, 2004). According to the 1951 Geneva Refugee Convention a refugee is defined as,

“A person who, owning to a well-founded fear of being persecuted for reasons of race, religion, nationality, membership of a particular social group or political opinion, is outside the country of his nationality and is unable or, owning to such fear, is unwilling to avail himself of the protection of that country” (UNHCR, 2003).

Upon fleeing, refugees find themselves reliant on local governments and NGOs to provide assistance with basic human needs. For over twenty years, Karen people have been fleeing Myanmar (Burma) to Thailand to escape persecution from the military junta which is in power. There are two million refugees living legally and illegally in Thailand along the Myanmar (Burma) border, including 50,000 in Mae La Temporary Shelter (TBBC, No Date; UNHCR, 2007).

Mae La has existed for over twenty years and has an improved water system. Drinking water is drawn from a nearby river and from springs, is chlorinated in most cases, and released into a pipe-tap network. Non-potable water is also available from various wells through out the camp, as well as surface water sources. This report summarizes the work completed by Harding (2008) and Rahimi (2008) on water distribution in the camp and by Vater (2008) on drinking water treatment in Mae La. Further detailed work is available in each of their theses.

1.1 Climate in Northwestern Thailand

The Tak region of northwestern Thailand is characterized by a tropical climate with wet and dry seasons (UN Thailand, 2006; ESS, 2002). The rainy season lasts from June to October, followed by a cool season until February. The weather turns hot and sunny between March and May (UN Thailand, 2006). This region of Thailand has an average temperature of 26ºC although there is significant variation during the year and, due to changes in elevation, over the region. Temperatures can range from 4ºC to 42ºC (Thailand Meteorological Department in ESS, 2002). The average annual rainfall in Mae Sot, Thailand
is 2100 mm (GOSIC, 2007). Figure 1-1 shows the monthly rainfall averages over the past 56 years. During the wet season there is a clear increase in precipitation, as more than 85% of the annual rainfall occurs during this period.

Figure 1-1: Average Monthly Rainfall for Mae Sot, Thailand (GOSIC, 2007).

1.2 Mae La Camp

The Mae La camp is a refuge for people seeking protection from the Myanmar (Burma) government. The camp is run by the United Nations High Commissioner on Refugees and has existed since 1984 (TBBC, No Date). Mae La is located near 16°30'N and 98°30'E in the northern region of Thailand about ten kilometers from the border with Myanmar (Burma) (Lumjuan, 1982; TBBC, No Date). The camp location is shown by the red circle in Figure 1-2. The nearest town, Mae Sot, is about 60 kilometers away from Mae La. The next nearest large city is Tak and Bangkok is about 500 kilometers south-east of Mae Sot (Google, 2007).
1.2.1 Environment
The Mae La camp is located in a valley surrounded by two ridges rising about 300 meters above the camp. These hills are distant extremities of the Himalayan mountain range which is mainly located northwest of Thailand. A river runs through the end of the two ridges and bounds the camp on the north. The Thai military protects the road that borders the camp and links it with the nearest Thai city of Mae Sot. These key boundaries are shown in Figure 1-3, where the camp is roughly circled in white. Figure 1-3 also shows some water infrastructure including several storage tanks and some spring locations.

1.2.2 Public Health and Water Supply
There are about eleven NGOs that provide services within the camp ranging from food and shelter provisions, health care, protection, and water supply. Aide Médicale Internationale (AMI) is currently charged with providing health care and water supply. The AMI Mae La Water and Logistics Coordinator oversees a staff of about 30 refugees who operate the water system on a daily basis. The water system consists of drinking water supplied to public tap stands from surface and groundwater sources. In addition, non-
potable water is provided throughout the camp by rope-pump and bore-hole groundwater supplies. The drinking water system is discussed in detail in Chapter 3.

Figure 1-3: Mae La Location, looking southwest (Google Earth, 2007; Lantagne, 2007).
2 WATER IN MAE LA

AMI maintains potable and non-potable public water supplies within Mae La. This chapter outlines the sources of water for the camp, the water system components, and the existing treatment processes for drinking water.

2.1 Drinking Water Sources

The drinking water sources in Mae La are surface water from the river and naturally flowing springs. The river runs east-west, cross-cutting through the north end of the southwest facing ridge that borders the camp. While the river is an important water supply for the camp, its existing quality and potential means of treatment are not the focus of this work. The thirteen springs that are used for public drinking water supply flow from the ridge that borders the camp to the southwest and their quality is the focus of this thesis.

There is annual variability in the volume of drinking water available in total and from each source. Figure 2-1 shows the available water volume for 2007 by month for river water, spring water, and total water. Flow volumes for April and July 2007 were not available and the values from 2006 are shown. Only the spring water volume was available for August 2007. The spring water sources become notably more important during and after the rainy season because their volume increases significantly. Relying on gravity-fed spring water during as many months as possible allows AMI to save on the cost of pump operation.

The dependence of available spring flow on rainfall is shown in Figure 2-2. Monthly spring flow in 2007 is shown as bars, while average rainfall in millimeters per month is shown by the black line. Rainfall increases during the wet season which runs from May to September. The spring flow increase has a lag of one to two months, as shown by volume increases in July through October. In particular Spring 10 is an important spring, contributing an annual average of 20% the total spring flow volume.
Figure 2-1: Division of 2007 Flow Volume from Storage Tanks by Source  
*Data from 2006, ** River Water Flow Rate Unavailable

Figure 2-2: Annual Rainfall and Spring Flow Averages (AMI, 2007; GOSIC, 2007).
2.2 Drinking Water System Layout

Drinking water in Mae La reaches over three-fourths of the population through public tap stands after passing through collection systems, pump stations, storage tanks, and distribution networks (Lantagne, 2007). The majority of the remaining fourth of the population gets their water from private sources. The water system was developed over time as the camp grew. This helps account for the many apparently separate systems, some of which were connected after their initial construction. The system is also complicated by the number of sources entering at various points. Some portions of the distribution system are supplied by one source while others are supplied by multiple sources, with annual variability in source ratio based on available water volume.

Water is pumped from the river or is gravity-fed from springs into storage tanks. There are five main tanks which supply the five largest distribution systems in the camp. The tanks are: A Tank, B Tank, C Tank, Christopher Tank, and MOI Tank. Several of the springs, including Springs 6 and 7, 10, 14, and 17, have their own storage tanks as well. The locations of some of these tanks, five of the springs, and the Spring 10 filter are shown in Figure 2-3. The orientation of the image is facing southwest toward the ridge, with the road running along the bottom edge of the image.

Spring 10 flows from its source over land into a collection basin and then into a collection structure called Box A. This concrete box contained large (10-25 cm) rocks which were replaced with 2-4 cm gravel in January 2008. From the collection structure the water is piped down a steep slope and passes through a vertical-flow rock filter (VFRF) that has rounded gravel media. After passing through this filter the water is again piped downhill to the HFRF. From the HFRF the water is piped to the SP-10 storage tanks and then enters the SP-10 distribution system and is connected to A Tank. Figure 2-4 is a schematic of the SP-10 collection infrastructure.
Most of the tanks, including the main ones listed above, are opened for distribution twice a day, generally for 3 hour periods from 6 to 9 AM and 3 to 6 PM. There is ample demand at the tap stands and people must wait in line to receive water. Typically, water is continuously collected throughout the distribution time and all available water is taken. Some of the smaller and isolated spring systems are always open as the spring water flows directly to tap stands.

Some private standpipes exist (such as those for the school or the hospital), but the vast majority of the tap stands shown in Figure 2-1 are public. It is estimated that tap stands provides the majority of the water supply to over three-fourths of the population (Lantagne, 2007). The water is free for residents of the camp.

There are three pumps used to drive the river water to tanks: Tim pump, Christopher pump and MOI pump. Tim pump brings river water to tanks A, B, C and Christopher; Christopher pump to both the Christopher and MOI tanks; and MOI pump to the MOI tank and recently, on an intermittent basis, to a storage pond located across the road from the camp.

A lower pumping rate occurs during the dry season because of the lack of available river water. Additionally, more water is available from the springs in August, so there is less need to pump water from the river.
2.3 Water Quality

AMI determines water quality in Mae La by turbidity measurement, microbial sampling, and chlorine residual testing. Information on turbidity is available for the past several years on a weekly basis at several locations throughout the distribution system.

Turbidity of the water supply is measured at a variety of locations throughout the distribution system—including spring collection boxes, storage tanks, and pump stations—normally on a weekly basis. Table 2-1 gives the monthly turbidity average at each of the springs. Values greater than ten NTU are highlighted. Spring 10 is the only spring to have turbidity consistently greater than 10 NTU. In particular, at Spring 10 the highest turbidity can be correlated with the rainy season. Figure 2-5 shows that the turbidity increases over the months of May to September.

Figure 2-5: 2007 Turbidity at Spring 10.
Table 2-1: 2007 Monthly Average Turbidity at Springs in NTU. Highlighted values exceed 10 NTU.

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2.4 Existing Water Treatment

Disinfection is the main form of water treatment to the river and spring water in Mae La. There is one filter in the system at Spring 10, the location of which is shown in Figure 3-2. As previously noted the main focus of this work is on the water from the springs, of which Spring 10 has been shown to have particularly high turbidity.

2.4.1 Disinfection

Currently, the only treatment for the river and spring water is disinfection by chlorination. Chlorine is a common disinfectant for treatment of water against disease-causing bacteria. According to Lantagne (2007), the distribution system had sufficient disinfection at the tap stands in August 2007.

Chlorination is more effective and lower dosages can be used in low turbidity waters (Sawyer et al., 2003). Since twelve of the thirteen springs maintain low turbidities throughout the year, chlorination of these sources should be effective throughout the year. The variation in the turbidity at Spring 10 requires that the water be treated before it is chlorinated in order to assure the effectiveness of the disinfectant.
2.4.2 Filtration

The only treatment other than disinfection as of January 2008 is a horizontal-flow roughing filter located between the collection and storage of Spring 10 water. The roughing filter has three compartments separated by internal walls constructed with off-set concrete blocks. Each compartment is 4 meters long, 1.5 meters wide, and filled with 0.8 m of filtration media. The media consists of loose gravel with a size range of two to four centimeters.

A pipe of three-inch diameter conveys water from the SP-10 collection box into the filter. This inflow pipe does not distribute the water over the operating width of the filter. The outflow pipe is a vertical perforated pipe that collects water over the depth of the filter and sends it to the SP-10 storage tank through an outlet located at the bottom of the HFRF.

The purpose of the filter is to maintain a consistent turbidity at SP-10, since there is an annual increase in turbidity during the rainy season. Figure 2-6 shows the turbidity increase at SP-10 collection Box A and the turbidity of the water at the SP-10 storage tank after passing through the HFRF. The filter achieved the treatment goal of less than 10 NTU on average for seven of the twelve months in 2007.

![Figure 2-6: 2007 SP-10 Turbidity at collection Box A and at storage tank. The HFRF operates between these two sampling points.](image-url)
2.5 Non-potable Water Sources

There are two main types of water access within the camp: drinking water and non-potable water. Drinking water is used for drinking and cooking, and hygienic water is used for bathing, laundry, hand and dish washing. This water is provided by public tap stands, while rope-pump wells, hand-dug wells or surface water serve as the sources of non-potable water. A series of deeper boreholes exist throughout the camp but are not currently used due to contamination and disrepair. These infrastructure points are visible in Figure 2-7.

People tend to store their water in containers on their porches and in their homes. If water goes unused, it is discarded and the containers are refilled the following day (Lantagne, personal communication, October 19, 2007). This makes understanding the actual water demand of the camp difficult, because not all the water collected is used.

Since it is not necessary that water for bathing, washing and other non-consumable water be disinfected through chlorination, there are a number of alternative access points throughout the camp. The primary alternate sources are the 63 rope-pump wells are located mainly at lower elevations in the camp. By UN definition it is an improved water source since there is a cover and concrete drainage area, but some of the wells are contaminated by sewerage (D. Lantagne, personal communication, October 19, 2007).

In order to collect water using rope-pump well, users place a container for collection beneath the opening of the blue PVC pipe and pull outwards on the pump’s metal handle. This mechanically drives water from a shallow ground water source to the surface and out
the blue pipe. In order to bathe, users will either collect water in a container to pour over themselves or place extremities at the opening of the pump one at a time to rinse off.

Some regions of the camp have very shallow ground water levels that can be accessed through hand-dug wells. These sources are generally discouraged as the open stagnant water is a breeding ground for disease carrying mosquitoes and the water is much more likely to be contaminated by sewage from nearby latrines.

Many people utilize the major river as well as a small stream that cuts through the camp a as sources for hygienic water. In the heart of the dry season, this stream can run dry and the river can run very low, decreasing or eliminating use.
3 MODELING OF THE WATER SUPPLY

The work by Rahimi (2008) focused on the modeling and mapping of the water supply system. It was motivated by the need to develop analysis tools to identify areas of improvements and provide a further understanding of the complexities of the system. AMI is currently planning an expansion in the geographical coverage of the water system, and a modeling tool can provide predictions of the performance of the system when new taps are added. The tool could also test alternative designs of the pipe system layout or the effect of increasing volume capacity.

One of the salient features of the water system is its intermittent nature: the distribution tanks are open twice a day for three hour periods from 6 to 9 AM and 3 to 6 PM. Typically, models assume a pressurized system which is not the case for intermittent supply. Analysis of the pipe volumes for the different subsystems and their corresponding distribution flow rates shows that most subsystems become pressurized in less than 10 minutes (Rahimi, 2008). This represents a small fraction of the distribution time, and hence, the assumption of pressurized flow is reasonable.

Another important feature of this system is that it is pressure driven: users will draw all the water available. In a typical continuous system, the demand dictates the system flow. This characteristic of the supply can be modeled using emitters instead of taps with associated demands: in the EPANET model emitters are devices such as nozzles or sprinklers that discharge to the atmosphere based on the pressure available.

3.1 Data Collection and Analysis

A field investigation was also conducted as part of the study of the water distribution system. The field work included surveying the coordinates of tap stands, valves, T-junctions and pipe-size reduction points using a GPS device. The size of these components and characteristics of tap stand use – public taps are typically used throughout distribution, while private taps and latrines do not have as much of a demand. We measured pipe lengths with a laser range finder and took notes on the system layout. With the help of the AMI staff, we also ran pressure, flow and salt calibration tests to verify the results of the model. Pressure and flow were directly measured at specific tap stands making sure to include samples from all modeled subsystems. Salt was added to the storage tanks such that concentrations of 80
percent of the U.S. Environmental Protection Agency (EPA) taste threshold. Samples were then collected at tap stands and salt concentration was analyzed over time using a conductivity meter. After the site visit, elevations were taken from a Digital Elevation Model (DEM) of the area provided by Bunlur Emaruchi from Mahidol University in Bangkok.

The computer program used for modeling is called EPANET: it is a free software code available from the U.S. EPA (Rossman, 2000) that simulates hydraulic and water quality behavior within pressurized pipe networks. Some of the key capabilities of EPANET include no size limitations on the network, time-varying controls (e.g. opening and closing valves), providing different types of mixing models to characterize mixing in storage tanks, and simulating pressure-driven nodes using the concept of emitters. To use the emitter capability as part of the model, we tested flow versus pressure of the type of tap used in the camp.

All coordinates, elevation and identifications of system components were entered into EPANET. Taps were included with their specific emitter characteristics; tanks were set to their corresponding size; pipe lengths and roughness as well as valve open or closed status were also entered. Controls were added to simulate the opening and closing of the tanks at specific times and to represent the salt test conducted. Finally, field calibration data was entered to allow for comparison with the calculated results of pressure, flow and salt concentration. The Mae La supply model in EPANET relies on the assumptions that all storage tanks started full, that only large-use public taps discharge significant quantities of water during distribution and that no water enters the tanks during distribution. This last assumption is reasonable because 75 percent of the water flow is pumped from the river outside of the distribution period and the portion of spring water entering tanks during distribution is not important within the three-hour distribution window.

3.2 Distribution System Model

The work discussed led to the construction of a computer model designed to serve as a tool for analysis and improvement of the drinking water supply system in Mae La. Figure 3-1 gives a global view of the Mae La water distribution system using the EPANET visual interface. Bunlur Emaruchi’s aerial photograph underlies the system. The river is situated at the northwest edge of the system and the ridge is to the southwest. The road to Myanmar can be observed crossing the picture from the southeast tip to the northwest with the system adjacent to it. The color coding of the map symbols in Figure 3-1 represents elevation and
can be read from the legend. Supplying tanks as well as secondary tanks receiving water are shown as rectangular symbols on the map. The many nodes (circular symbols) represent tap stands, joints and tees. Pipelines are shown as blue lines.

![Figure 3-1. Overall view of water distribution system in EPANET](image)

The skeleton of the system is shown in Figure 3-2 in the same orientation as Figure 3-1. The different subsystems within the overall distribution system are shown. Spring 6/7 and Spring 8 are sub-components of the C Tank subsystem since C Tank can supply water to their secondary tanks. The TB Village is a sub-component of the Spring 17 subsystem since water is pumped from Tank 17 to the TB tank. As can be observed, some subsystems can potentially interconnect but the valves separating them are usually left closed: these valves are illustrated by black circles in Figure 3-2. EPANET can calculate pressures, demand, and water quality at different nodes as well as flows, velocities and head loss in pipes throughout the distribution period.
Pressures, flows and salt concentrations were measured in the field in order to compare them with the results of the distribution model. Pressure calibration revealed a coefficient of determination ($R^2$) of 0.36. The major source of error resides in the elevation of tap stands. The GPS coordinates recorded have a horizontal (X-Y) accuracy of about 15 meters. This magnitude of error translates, depending on the slope, to elevation errors from 3 to up to 15 meters when using the DEM. This is commensurate with the error recorded between measured and calculated pressure results. In addition, this error in pressure was random including over-estimates and under-estimates at all ranges, with no particular tendency by subsystem, confirming that it is not an error in methodology. Since flow is directly related to pressure, a similar level of error was recorded between calculated and measured flow rates with an $R^2$ value of 0.38.

Salt concentrations on the other hand do not directly depend on elevations: they are predominantly a function of the mixing at storage tanks. The $R^2$ between calculated and measured results from the salt-tracer test is much higher than that for flow and pressure calibration: it amounts to 0.82. Measured concentrations revealed that salt did not completely mix in the storage tanks but short-circuited them to an extent. A two-compartment mixing model was adopted to simulate such dynamics: it includes a bottom
compartment, the active compartment with inflow and outflow of salt, and a top “dead-zone” compartment. It was found that a fraction of 12 percent for the active compartment best reflects the reality of the system. Salt analysis also showed that dispersion of matter throughout the system is not an important process. In addition, it was found that some of the flows of the system did not follow the expected controls, and that some flow actually does enter certain storage tanks during the distribution period.

### 3.3 Analysis of Distribution System

The distribution system model was used to evaluate three alternative scenarios to improve system performance. The objective of the first and second scenario is to increase the flow rate at taps of low supply; the third scenario aims at adding taps to parts of the camp without easy access to running water. The first scenario consists in opening the valves shown in Figure 3-2 to connect subsystems. Figure 3-3 shows the evolution of flow distribution between the original and modified system. Under the modified system, the Spring 14 tank drains after an hour of distribution, and Figure 3-3 shows the flow distribution both before and after drainage.

![Flow Distribution Graph](image)

**Figure 3-3. Numerical distribution of flow from opening system connecting valves**

During the first hour of distribution after opening the valves (blue line in Figure 3-3), the decrease in percentage of low-flow taps is significant in the 16 to 22 liters per minute range by pushing 3 percent of them to higher flow rates. The bulk of the benefit however occurs...
for higher flow taps between 33 and 47 liters per minute by decreasing their numbers by 6 percent. The downfall of this scenario occurs after the Spring 14 tank drain, as illustrated by the green line in Figure 3-3: there are 2 percent more taps with zero flow. Hence, this scenario benefits taps of large supply more so than taps of low supply and then completely drains parts of the water supply. This scenario is not recommended.

The second scenario consists of adding connecting pipes between subsystems of high pressure and those of low pressure. Figure 3-4 shows the evolution of flow distribution between the original and modified scenario. Even though the total number of taps under 20 liters per minute is still the same (blue and red lines join), there is an improvement in the lowest flow rates: 2 percent of the taps are no longer under 10 liters per minute and 2 percent as well are no longer under 15 liters per minute. These 2 percent have now moved to the 15 to 25 liters per minute range, a more reasonable flow rate. Also, taps with flows between 25 and 28 liters per minute have decreased by 2 percent. This scenario is recommended because it would increase the flow rate of low- and medium-supply taps and it involves a low investment and no added water supply.

Figure 3-4. Numerical flow distribution under the adding pipe connections scenario

For the third scenario, taps were added in areas of the camp without easy access to water: these were defined by using the results of mapping the current system and the
population distribution by Harding (2008). Figure 3-5 shows the geographical evolution of flow distribution between the original and modified scenario with the new taps illustrated by white stars. It shows that reasonable flow rates can come out of all the new taps. Comparing the original scenario against that with new tap stands, one can see that the latter covers a larger geographical area of the camp: it has a denser southeastern distribution as well as northwestern. Hence, these new taps were successful in providing water to these areas and this scenario is recommended.

An additional recommendation for increasing the water supply in the camp was found from analyzing tank levels during the distribution period. Figure 7 shows the evolution of water level in storage tanks through a three-hour distribution period. The three tanks that drain during the distribution period (Spring 6/7, Spring 8 and TB) only make up about 5 percent of the total volume of water supplied. Figure 3-6 shows that none of the major tanks drain fully after 3 hours of distribution. The limiting tank with the most use of its capacity is the Christopher Tank which drains about 53 percent of its volume. In order to make full use of the capacity of tanks, a safe recommendation is to increase the period of water supply from 3 hours to 4½ hours.
3.4 Recommendations

Further work would be to refine the model by obtaining more accurate elevation measurements, as elevation error and its relation to pressure and flow is the main weakness of the simulation. Accurate elevations should be obtained for taps and tanks. Particularly, it would be important to obtain precise elevations of taps close to tanks because these taps have low pressures and are located on steep slopes. To complement these elevation measurements, more calibration tests should be run at the taps surveyed and throughout the system.

To make a more complete model of the system, the flow of springs into tanks could be modeled as well as its seasonal variation. Also, after re-defining the layout of the system to meet goals of demand per household in the whole area of the camp, the next step would be to run an optimization model: it would vary pump and tank capacity, and take into account budget and river flow constraints to lead to an era of extended, accessible, drinking water supply.


4 GIS REPRESENTATION AND ASSESSMENT

A geographic information system (GIS) is a useful tool to understand spatial relationships and visualize problems in new ways. This work utilizes a GIS in coordination with a computer model created by Navid Rahimi (2008) to better understand the condition of water supply within Mae La camp, Thailand.

This assessment distinguishes between two main types of water access within the camp: consumable water and hygienic water. Consumable water is disinfected with chlorine and used for drinking and cooking, and hygienic water is used for bathing, laundry, hand and dish washing. Consumable water is provided by public tap stands, while rope-pump wells, hand-dug wells or surface water serve as the sources of hygienic water.

4.1 Data Collection and Analysis

Various data sources were needed to complete this work. Global Positioning System (GPS) data of major water system infrastructure points were collected during a site-visit. Location data for points not visited were received from Daniele Lantagne; 130 of the 152 tap stands were visited during the site-visit. Additionally, Dr. Bunlur Emaruchi from the Faculty of Civil Engineering of Mahidol University in Bangkok supplied a DEM which was received during the site visit. Upon return to Cambridge, home location data were collected through inspection of Google Earth images.

With respect to the inspection of home locations, since the highly populated areas, such as in the northeast section of camp, have sparse vegetation cover, it is likely that more homes were unidentified in the less populated areas. This may affect the results since the less populated areas also tend to be further away from infrastructure points of interest such as tap stands and rope pump wells. It is thus possible that the results are skewed so that a fewer number of homes, both as a percentage and a raw number, are identified as being undesirably far from water points.

4.2 Results

Through the analysis of home, tap stand, and rope-pump well locations along with outputs from the EPANET model, the effectiveness of water access within Mae La camp is
accessed. The work identifies homes and regions with inadequate service concerning one or more of the following:

1. Located further than 115 meters to nearest tap stand
2. Located further than 180 meters to nearest rope-pump well
3. Availability of less than 50 liters per home per day

4.2.1 Tap Stand Proximity

It has been shown that the amount of time needed to collect water (round-trip) correlates strongly with consumption, and when the return trip travel time to source water is less than about three minutes, water consumption drastically increases (WELL, 1998). This analysis assumes that each home gathers drinking water from the nearest public tap stand. The range of comfortable walking pace considered was 75-85 meters per minute (Bohannon, 1997). To be conservative speed of 75 meters per minute is used which equates to a maximum allowable tap stand distance of 115 meters.

Figure 4-1 shows an overall view of the camp with homes represented by different colors based on distance to the nearest viable tap stand. Tap stands are considered viable if public drinking water is provided for collection. For example, public latrines and private taps for NGOs are not included.

There are 349 of the 7,117 homes identified (less than 5%) located further than 115 meters from the nearest viable tap stand.

Figure 4-2, a histogram of the results, includes the Spring 2 homes. When these homes, which are well supplied by natural springs, are not included, the number of homes with tap stands located more than 200 meters away is reduced by 60% (Terville, personal communication, January 2008). Fifty percent of homes are located between 30 and 60 meters from a drinking water source.
Figure 4-1: Home Distance to Nearest Tap Stand.
From Figure 4-1, we see a large cluster of homes of concern located between the Spring 17 and A systems (“Low Coverage Region”) in addition to the Spring 2 region. Besides these two major regions, homes of concern are sparingly distributed mostly along the mountain ridge that runs along the camp border furthest from the access road. Placing taps along this ridge is difficult as the slope becomes very steep and many of the homes are located at higher elevations than the system storage tanks. Since the systems are run by gravity, it is impossible to supply tap stands at these elevations. Improvement appears possible for the “Low Coverage Region”. This is a large cluster of homes and the elevations are not prohibitively high in comparison to the A and Spring 17 tanks.

4.2.2 Rope-Pump Well Proximity

It is customary for people to bathe at the rope-pump wells and bring their laundry to the well to wash near the water. This way, large amounts of water do not need to be carried back to the home and use mainly occurs at the well. On the other hand, each member of the household must travel to the well to bathe while one person can carry drinking water for all. As a result, the critical distance limit is set at 180 meters or approximately one and a half times the critical tap stand distance. This criterion is used in Figure 4-3 to identify homes that are a problematically distant from a rope-pump.
Under this criterion, just over 1000 homes, or 14%, are an unreasonable distance from the nearest rope-pump well (Figure 4-4). Over one half of the homes have a rope-pump well somewhere between 30 and 100 meters away. There are a much greater number of homes located far from rope-pump wells compared to tap-stands, but this may not be easily remedied and there are additional sources of washing water.

It is likely that drilling wells along the mountain ridge is not economically feasible given the greater depth to the water table from the increased elevations. Since a rope-pump well relies on the ability of the user to pull water from the water table to the surface, the wells are ill suited for locations where this distance is large. Also, there is a stream that runs west through the camp to the river in the northwest which can act as an alternative supply for some areas.
Figure 4-3: Home Distance to Nearest Rope-Pump Well.
4.3 Volume of Water Per Household

Through linking the results of Navid Rahimi’s EPANET model (Rahimi, 2008) with home locations, an estimate of available water volume per home is made. Rahimi’s model predicts the average flow rate for 102 of the 139 viable tap stands (some systems were not included in the EPANET model). We use homes as a proxy for population, and assuming an even distribution of a population of 45,000 among an estimated 8,500 homes, there would be between five and six people per home (F. Pascal, personal communication, April 21, 2008). Considering a conservative estimate for per capita water consumption, 50 liters per home per day is the designated minimum volume of consumable water desired.

Predicted flow rates were multiplied by the six-hour distribution time to get flow volume. For each home, the flow volume for the nearest tap stand was divided by the total number of homes associated with that tap-stand. This is daily availability of drinking water per home and shown in Figure 4-5.
Figure 4-5: Daily Home Water Availability.
The homes of concern, shown in red and orange, are scattered throughout the camp. There is no single subsystem within the overall network where flow is low and no geographic similarities between the homes of concern, such as being located along the steep mountain ridge. Homes for which the closest viable tap stand was not included in the EPANET model are shown in black.

Figure 4-6: Water Volume Distribution - Histogram.

A total of 809 homes, or 15% of those considered, are categorized as unable to obtain 50 liters of water per day (Figure 4-6). By tracing these underserviced homes back to the originating taps, we find that there are 15 tap stands of concern.

There is definitely error in the model results because the model predicts flow rates of zero liters per minute at nine of the tap stands. Flow was observed at these tap stands during the site visit, however. It is most likely that these errors are related to the elevation assigned to the tap stands based on the GPS location. As discussed by Harding (2008, Section 4.3), the GPS location error can create significant error in elevation. Since the model is driven in large part by these elevation differences, the model results are sensitive to these errors (Rahimi, 2008). Excluding the nine tap stands with zero flow, 365 homes, or 7%, are unable to collect sufficient water volume.
4.4 Conclusion and Recommendations

Overall this research shows that the vast majority of residents in Mae La have sufficient access to water. A water use survey is recommended in order to verify the findings of this research and modify the GIS tool for future work. The assumptions that every home utilizes the rope-pump well or tap stand that is of closest proximity may or may not be valid. A major area of concern, especially regarding the EPANET model results, is in attaining accurate locations and especially elevations of tap stands and water infrastructure points within the camp.

4.4.1 Overall Water Access

This research used GIS to assess three major indicators — home distance to tap stands, home distance to rope-pump wells, and volume of drinking water per home — with results summarized in Table 4-1. The overall results show that the access issue of least concern is proximity to public tap stands.

Table 4-1: Summary of Homes with Inadequate Access.

<table>
<thead>
<tr>
<th>Homes with Far Taps*</th>
<th>349</th>
<th>(5% of 7,117)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Homes with Far Rope-Pump Wells</td>
<td>1,017</td>
<td>(14% of 7,117)</td>
</tr>
<tr>
<td>Homes with Low Volume</td>
<td>809</td>
<td>(15% of 5,500)</td>
</tr>
</tbody>
</table>

*Reduces to 210 (3%) when not including Spring 2 region

There are homes which fail more than one test, however. Table 4-2 shows a breakdown of the results considering that some homes will have multiple problems. Of homes identified, 73% are adequately serviced. Roughly one fifth of these homes are located nearest to tap stands not included in the EPANET model and therefore the volume test was not completed.

4.4.2 Potential Improvements

There are a variety of concerns regarding these results and what service is actually provided in the camp. The proximity to the nearest rope-pump well may not relate directly to water use since there are additional sources for non-drinking water such as bore holes and surface water. A water use survey that gathers information from a variety of homes dispersed throughout the camp would help better understand the extent of these alternative sources.
The survey should account for seasonal change either by clearly asking questions about the different season or by surveying at multiple points throughout the year.

Table 4-2: Breakdown and Overlapping Burdens for Home Water Access.

<table>
<thead>
<tr>
<th></th>
<th>Flow Data</th>
<th>No Flow Data</th>
<th>Total</th>
</tr>
</thead>
<tbody>
<tr>
<td>Far Tap, Far Well &amp; Low Volume</td>
<td>18</td>
<td>-</td>
<td>18</td>
</tr>
<tr>
<td>Far Tap &amp; Low Volume</td>
<td>18</td>
<td>-</td>
<td>18</td>
</tr>
<tr>
<td>Far Well &amp; Low Volume</td>
<td>52</td>
<td>-</td>
<td>52</td>
</tr>
<tr>
<td>Far Tap &amp; Far Well</td>
<td>78</td>
<td>93</td>
<td>171</td>
</tr>
<tr>
<td>Far Tap Only</td>
<td>37</td>
<td>105</td>
<td>142</td>
</tr>
<tr>
<td>Far Well Only</td>
<td>471</td>
<td>305</td>
<td>776</td>
</tr>
<tr>
<td>Low Volume Only</td>
<td>721</td>
<td>-</td>
<td>721</td>
</tr>
<tr>
<td>Near Tap, Near Well &amp; High Volume</td>
<td>4,105</td>
<td>1,114</td>
<td>5,219</td>
</tr>
<tr>
<td><strong>Total</strong></td>
<td><strong>5,500</strong></td>
<td><strong>1,617</strong></td>
<td><strong>7,117</strong></td>
</tr>
</tbody>
</table>

This survey could strive to understand how different groups, based on geography, wealth, ethnicity, gender, or age, access and utilize water. While logically homes located in the very steep sections of camp far from a public tap may adapt to using less water, there may be other subtle differences about the use of bore holes based on age or gender. The survey should ask which tap stands are frequented by the home. Do different members of the home prefer different tap stands and what are the perceived benefits? It was observed during the field visit that some systems (B System, for one) had perceivably higher pressures which resulted in shorter lines at the tap stands. How much further is a person willing to walk in order to avoid waiting for water?
5 RECOMMENDATIONS FOR SPRING WATER QUALITY IMPROVEMENT

Spring water is an important water source for Mae La. It accounted for an average of 50% of the water supply in 2007 and climbed to 80% of the water supply during the rainy season. The spring water in Mae La presents several water quality issues because of the number of springs and the annual water quality cycle. This chapter describes the design recommendations for spring water treatment based on the work of Vater (2008). Her work contains detailed theoretical background of horizontal-flow roughing filtration.

5.1 Design Parameters

Water treatment design is based on influent water quality and treatment objectives. Based on the existing information about spring water quality in Mae La, four main design parameters have been identified. Of these, turbidity is especially important because the effectiveness of disinfection by chlorination is lowered at higher turbidity (AWWA, 2003; Wegelin et al., 1991). Water from the largest volume springs is chlorinated before release into the distribution system. The bacterial load of the water is the second important parameter because of the potential for disease-causing organisms in the water supply, particularly in the sources that are not chlorinated. The last two design parameters have to do with the size and location of the spring, both of which are unique to Mae La due to seasonal flow variation and steep land slopes.

5.2 Pre-treatment

As described above, the main goals of treatment are to remove turbidity and bacteriological contamination and to normalize the quality throughout the year. Due to the range of turbidity seen in Table 2-1 and Figure 2-5, some sort of pre-treatment is recommended (Okun and Schulz, 1984). Wagner and Lanoix (1959) write that the simplest technologies are the best because of low maintenance requirements. This is particularly important in Mae La, where the availability and turnover in maintenance staff predicates low maintenance.

Pre-treatment options include several low-technology processes that improve water quality without significant labor or mechanical investment (Wagner and Lanoix, 1959). The
influent water quality and quantity dictate the type of pre-treatment. Horizontal-flow roughing filtration is the selected pre-treatment technology for use in Mae La. Other treatment technologies are outlined by Vater (2008).

5.3 Field Work

Information on the condition of the existing spring collection and treatment infrastructure was gathered on a site visit to Mae La Temporary Shelter in January 2008. Water turbidity at Spring 10 (SP-10) was measured several times at various locations. Flow rate tests were conducted at the SP-10 filter.

5.3.1 Turbidity Sampling

Turbidity was used to demonstrate the effectiveness of various roughing filters along the pipe that runs from SP-10 Box A to SP-10 storage tanks. There are two rock vertical-flow roughing filters (VFRFs) which precede the main SP-10 horizontal-flow roughing filter (HFRF). The small VFRFs have larger-diameter filter material than the HFRF.

The short length and large media in Box A and the VFRF predict that neither is effective at removing turbidity from the water. Turbidity was sampled twice at multiple points along the collection pipe. The first set of data was taken before the HFRF was cleaned, while the second set was collected after cleaning. Both sets are under natural environmental conditions for January, which falls in the dry season. Stream 1 and Stream 2 are two sample points in the SP-10. Data are available from Vater (2008).

The turbidity in the stream and collection basin is between 30 and 40 NTU, with one spike to above 70 NTU. From Box A through the HFRF inflow, the turbidity is again between 30 and 40 NTU, but usually nearer to 30 NTU. The HFRF is the only treatment process that removes a significant amount of turbidity. In one case, when the filter was not cleaned and had not been modified, more than 50% removal was achieved. In another case, when the filter had been cleaned and the baffles added, more than 85% removal was achieved.

Two conclusions can be drawn from these results. First, neither Box A nor the VFRF is treating the water. The treatment is all achieved in the HFRF. Second, the clean, modified filter is more effective than the dirty, unaltered HFRF.
5.3.2 Filter Flow Tests

The SP-10 roughing filter consists of three compartments each with the same filter material. Upon first inspection, the free surface of the water was above the filter media in the first and third compartments. The filter media in the second compartment had a layer of sediment on top of it. The construction of the filter, without inlet or outlet weirs, suggested that short circuiting could be occurring. Figure 5-1 shows the plan layout of the filter, including the walls separating the three compartments. These walls are made of offset cement blocks. The theoretical flow path is shown with dashed arrows. Potential short circuits are shown by the arrows. The inflow is a pipe through which water free-falls onto the filter media and the outflow is a vertical pipe which is perforated over depth. To determine the residence time of the filter under different conditions, three tracer flow tests were conducted under varying filter and influent water conditions. The first test was completed before the filter was cleaned or modified. The next two tests were completed after the filter was cleaned and the baffles were added.

The filter was cleaned during the second week of field work. In addition, gravel was added to the first and third compartments so all the water now flows below the surface of the medium. While the gravel was removed from the filter, two baffles were added in order to direct flow through the entire filter volume. Each baffle consisted of a plastic sheet placed over a portion of the cinder-block interior wall. The layout of these baffles is shown in Figure 5-2.

The results of the tracer tests are divided into two sections, one for the two tests run with dry season turbidity and the other for the test run with simulated-rainy season turbidity. The results demonstrate that the baffles increased the residence time of water in the filter by approximately three times and that the clean HFRF is capable of greatly reducing even high turbidity loads.
5.3.2.1 Dry Season Load

The mean residence time of the tracer in the first test was about 30 minutes. This test established a background to which the post-baffle installation tracer test could be compared. Although the exact relationship between the flow velocity of the tracer and the
water is not known, such a low tracer residence time likely indicates short-circuiting of the filter. This means that the treatment potential of the filter was not maximized.

The second test results showed the residence time of the tracer to be about 95 minutes. This test was conducted after the baffles were installed, demonstrating that this simple installation increased the functionality of the HFRF. The relative residence time of water in the filter increased three times after installation of the baffles. Figure 5-3 shows the data for both the tests. The residence time for the tracer found in these studies is not the residence time of the filter because of the density difference between the tracer and the water. The tracer residence time is about 20 minutes longer than the theoretical residence time of water in the filter.

![Figure 5-3: Dry season flow test results. Time shown is for the tracer.](image)

In conclusion, for these two flow tests under similar influent flow volume and water quality, the baffles increase residence time and produce a higher quality effluent. The turbidity in the first case was reduced 53% from 30 NTU to 14 NTU. After adding the baffles the reduction was 86% from 38 NTU to 6 NTU. Part of this improvement can be
attributed to cleaning the gravel. The clean HFRF with baffles can achieve the treatment goals for SP-10 turbidity during the dry season.

5.3.2.2 Rainy Season Load

The order to better understand the performance of the HFRF during rainy season turbidity levels, which range from 200 to 400 NTU, high turbidity water was sent through the SP-10 HFRF. The water was agitated at the collection basin, raising the turbidity to between 100 and 500 NTU, or three to 15 times the dry season average turbidity of 30 to 40 NTU.

The HFRF removed over 90% of the turbidity over the duration of this test. Even still, because the influent turbidity was so high, effluent turbidity was 20 to 30 NTU. This is above the treatment goal of the HFRF. The SP-10 treatment processes, including the HFRF, could be modified so that even during the high turbidities of the rainy season the goal of water with 10 NTU or less can be met.

During this high turbidity test another tracer test was also conducted. It resulted in a tracer residence time of about 90 min. This result is similar to the result of the other test run after the baffles were installed.

5.3.3 Results

Based on the SP-10 Filter-specific filter coefficients summarized by Vater (2008), it is possible to predict performance of a horizontal-flow roughing filter design for SP-10. If the existing filter is lengthened, or a second filter is built, filter performance will improve.

The existing filter is 12 meters in length. Only under the Dry Season, Clean Filter coefficient condition, an effluent turbidity of less than 10 NTU is achieved. Notably, when the filter length is 24 m all the flow scenarios except Wet Season, Dirty Filter, predict an effluent turbidity of less than 10 NTU. Not until the filter is 38 m long, more than three times the length of the existing filter, do all four flow conditions achieve below 10 NTU effluent values. The Wet Season, Dirty Filter coefficient is also a liberal estimate and so it could take an even longer filter to achieve the desired treatment.

Ideally the Wet Season, Dirty Filter flow condition would not occur. This means that a filter of double the existing length should achieve the desired effluent turbidity level during the remaining possible conditions. The main limitations on a long filter are
construction and maintenance costs and land availability. At the site of the existing SP-10 Filter, there is not enough land area to triple the filter length. There is enough space to double the filter size.

5.4 Recommendations

Vater (2008) recommends treatment technologies that are easily implemented and operated to improve Mae La spring water quality. First, modifications, including maintenance, to the SP-10 filter that will improve its performance are addressed. Second, the application of a general filter design for the other twelve springs is discussed.

Information about the spring water quality and treatment was gained through the analysis of turbidity and flow volume from 2007 and field study of the SP-10 horizontal-flow roughing filter (HFRF). The combination of an additional filter and improvements to the existing filter will meet the treatment demands of the system, particularly that which is a problem at present—high turbidity during the rainy season.

5.4.1 SP-10 Box A and VFRF

Based on the turbidity measurements taken along the water-collection structures of SP-10, both the VFRF rock filters preceding the HFRF were shown to be ineffective. Based on this information the VFRF was removed and the rocks in Box A were replaced with 2-4 cm gravel. The effectiveness of such a short, 1 m, HFRF in Box A is unknown, but the use of gravel instead of large rocks should at least reduce the velocity of flow through Box A, preventing re-suspension of solids from the bottom of the box.

5.4.2 SP-10 Filter Modifications

The existing HFRF filter has neither an inlet weir, an inflow distribution chamber, an outflow collection chamber, nor an outlet weir. Weirs are used in HFRFs for maintaining constant water volumes within the filter by only allowing inflow and outflow from above the top of the filter media. Chambers located at the inlet and outlet distribute and collect, respectively, the flow over the depth of the filter (Wegelin et al., 1991).

Since the saline tracer was not passive, it cannot be used to determine vertical distribution of flow at the inflow pipe. The tracer studies and baffle addition did show that there was horizontal short-circuiting in the filter which was corrected by the baffle
installation. The installation of a perforated pipe to distribute the flow vertically at the inlet is recommended. The pipe should be of 3-inch or 4-inch diameter and perforated every 0.5 to 1-inch. It should be installed at the inflow of the filter where the influent water will flow into it. This is a simple improvement that will improve vertical distribution of the influent water.

The combination of an outflow collection chamber and an outlet pipe located at the top of the filter can be used to collect water over the depth of the filter and maintain a constant volumetric flow through the filter (Wegelin et al., 1991). The existing outlet pipe in the SP-10 filter is a vertical-slotted pipe with an outflow at the bottom of the filter. This construction does not hold a constant volume of water in the filter, allowing water to exit the filter more quickly than the theoretical residence time of the filter. In order to achieve maximum filter performance, water should remain in the filter for as close to the theoretical residence time as possible.

If, instead, the effluent water was collected at the top of the filter, then the flow velocity through the tank would be regulated and a constant volume of water would be held in the filter. This can be achieved by using an existing pipe located at the top of the filter and by collecting the water over the depth of the filter in a collection chamber. The construction of the chamber will require a slight reduction in the length of the filter, but will allow water to flow easily into the existing outflow pipe. To make the chamber, a slotted wall should be constructed 0.25 m away from the existing wall which contains the effluent pipes. The wall would be built with off-set concrete blocks in the same construction style as the internal walls of the filter. Design drawings can be seen in Vater (2008).

The filter already has an outflow pipe near the top of the filter. This outflow pipe would need to be connected on the outside of the filter to the existing pipe which flows to the SP-10 ring tanks. The new pipe connection should be the same diameter as the existing outflow pipe so that there is no loss of head as the water enters the pipe to the SP-10 ring tanks.

5.4.3 Second equivalent filter

The second recommendation in order to meet the treatment goals of the system is to build a second filter at SP-10 which is equivalent to the first, including the modifications discussed above. Doubling the filter length to 24 m means the treatment goal of less than
ten NTU in the effluent water is achieved in three of the four flow scenarios. The only flow scenario in which it is not achieved is the Wet Season, Dirty Filter scenario. Ideally this condition will not occur with proper filter maintenance.

The second filter should be operated in parallel with the existing filter. This means the influent pipe should split and half the flow should travel to the first filter and the other half should travel to the second filter. A valve should be installed in each pipe just downstream of the point in the pipe where the flow is divided in order for all the flow to be diverted to one of the filters if necessary. The construction of outflow collection chambers in both of the filters will ensure that the flow entering either filter is treated for the same amount of time.

5.4.4 Maintenance

In order to keep the new and existing HFRFs operating so they meet the treatment goal, both filters need to be maintained. The first step to maintaining the filters is daily or weekly monitoring of influent and effluent turbidity levels in order to determine the efficiency of the filter. To collect useful turbidity information, a regularly calibrated electronic nephelometer should be used. The LaMotte turbidimeter currently owned by AMI would be effective if regularly calibrated. In addition the staff members that are using the turbidimeter should be trained to use it and collect accurate data.

Second, the filters will need to be cleaned when the effluent water quality is not meeting the treatment goals. A benefit of the two parallel filters is that one filter can easily be taken off-line for cleaning by closing the valve in the inflow pipe. In addition, during lower volumes of flow during the dry season, it may be possible to only operate one of the filters, thus reducing the labor demand of cleaning the filters. Another benefit of having two filters is that the total amount of solids collected will be distributed over two filters instead of one, thus theoretically increasing the amount of time between cleanings.

Currently the filters are cleaned manually. Mechanical cleaning of HFRF is an option which would reduce the need for labor and reduce the amount of time the filter is offline. In order to effectively remove solids from the filter, turbulent flows of 0.5-1.5 meters per minute are needed. Wegelin (1996) writes the ideal flushing flow would be 1.0-
1.5 m min\(^{-1}\). Taking 1.0 m min\(^{-1}\) as the flushing flow goal, the volume of water that needs to be flushed is 18 m\(^3\) min\(^{-1}\).

The water will flow through these pipes at 4 m s\(^{-1}\). From the flush rate and the velocity of water through the drainage pipes, the area of pipes is calculated as five times the present area. Five three-inch pipes need to be added to each compartment to achieve this. With six three-inch drainage pipes per compartment, the flushing velocity of 1.0 m min\(^{-1}\) is achieved. The amount of time it will take for the filter to drain under these conditions is 45 seconds.

The drainage velocity assumes that the height of water in the filter remains constant over the flushing. Since the flushing volume is the same as the volume of the filter, 15 m\(^3\), it will be necessary to replace this volume at the same rate at which water exits the filter in order to maintain the same head over the duration of flushing. First, this means that this volume of water must be available. AMI is planning on constructing storage tanks for this purpose along the SP-10 collection pipeline, at the location of the VFRF. Second, this volume of water must be able to reach the HFRF at the same rate that water is being flushed out.

In order to achieve this flow rate the head in these tanks needs to be maintained as well. This requires storing two times the necessary volume or 30 m\(^3\). This can be achieved in 12 eight-ring tanks made with one-meter diameter rings. The ring tanks should be constructed in two rows of six. Six of the tanks will have two three-inch pipes which connect to the HFRF. The other six tanks, placed behind the first row, will each have two three-inch pipes which connect to one tank in the first row.

This flushing system is a significant investment of capital. It will require the addition of twelve ring tanks and the pipes necessary to connect the tanks to the HFRF. If AMI chooses to construct this system, I recommend they make the pipe connections between the storage tanks and the HFRF temporary. This will require that the pipes are reconnected for each flushing, but this should only be one day in every two to four months. During the months they are not in use they can be locked at A Tank or the Mae La 2 Office. Detaching these pipes will require labor to set up on the day of the flushing, but since the flushing will take less than one minute, the entire process should only require the services of day laborers for one day. This is in contrast to the current manual medium regeneration method which requires multiple day laborers working for five days. It should be reiterated that manual
medium regeneration is an effective means of cleaning the HFRF and is as recommendable as mechanical regeneration.

5.4.5 General filter design

One purpose of this work was to provide a general filter design that can be constructed at the twelve springs currently without filters. Review of the turbidity levels collected weekly in 2007 shows that no spring other than SP-10 has elevated turbidity throughout the rainy season.

The springs which had a monthly average turbidity over 10 NTU are SP-2, SP-8, and SP-10. SP-2 and SP-8 had only one month with an average exceedance, while SP-10 exceeded 10 NTU every month. In addition to these springs, SP-7 and SP-12 each had at least one weekly exceedance of 10 NTU, but these did not result in monthly averages more than 10 NTU. Since only SP-10 has consistently elevated turbidity, it is not recommendable to construct a roughing filter at any of the other twelve springs. These data are shown in Table 2-1.
REFERENCES


